Mesoscale and Hybrid Models of Fluid Flow and Solute Transport

Yashar Mehmani

Department of Energy Resources Engineering Stanford University, 367 Panama St. Green Earth Sciences 050, Stanford, California 94305-4007, USA

ymehmani@stanford.edu

Matthew T. Balhoff

Department of Petroleum and Geosystems Engineering The University of Texas at Austin, 200 E. Dean Keeton St. Stop C0300, Austin, Texas 78712-1585, USA

INTRODUCTION

Fluid flow and reactive transport is relevant to many subsurface applications including CO₂ sequestration, miscible/immiscible displacements in enhanced oil recovery, wellbore acidization, pollutant transport, and leakage/remediation of nuclear waste repositories. In all these scenarios, one or more fluid phases flow through the complicated geometry of the pore space, while advecting one or more chemical species along their flow streamlines. Simultaneously, the chemical species undergo molecular diffusion, due to their Brownian motion, allowing them to randomly jump from one streamline to the next. In the case of fluid–fluid or fluid–mineral reactions, chemical species may be transformed, potentially leading to precipitation and/or dissolution of solid minerals that alter the geometry/topology of the pore space. This in turn affects the velocity field of flow, and thus transport via advection/diffusion. Such complicated feedback between these pore-scale processes could give rise to "emergent" manifestations at larger scales. These manifestations are referred to as "emergent" because they cannot be foreseen from the behavior of the individual pore-scale mechanisms involved. In order to make reliable predictions of flow and transport at any scale of interest, accurate models need to be developed.

Two spatial scales are commonly identified with a porous medium: the "micro/pore scale" (1-100 µm) and the "macro/continuum scale" (>1 m). The former is the fundamental scale in which physical processes (flow, transport, and geochemistry) take place, and the porous medium is regarded as discrete in nature (void space vs. grain space). The latter is a more practical scale, where we would ultimately like to have a reliable description of flow and reactive transport, and the porous medium is regarded as a continuum. The macroscopic parameters appearing in the description of continuum models, such as permeability or dispersion coefficient, are typically extracted from experiments or stand-alone pore-scale simulations. While such a "hierarchical" upscaling approach is often appropriate, there are many practical problems for which this is not the case. In these problems, perturbations of state variables (e.g., concentration) at the pore scale are tightly coupled to their averaged quantities at the continuum scale. In other words, a clear separation between the temporal scales and/or the spatial scales does not exist. An often encountered scenario is solute transport in the presence of strong fluid-mineral reactions (Kechagia et al. 2002; Battiato and Tartakovsky 2011). Any improvement to the continuum description of these processes requires a fundamental understanding of the relevant physics at the micro scale. In this effort, pore-scale models are an essential asset.

This chapter is divided into two parts. In the first, we review various pore-scale modeling approaches for simulating flow and reactive transport, and discuss their advantages and limitations. We focus primarily on single-phase flow and transport, and give a brief discussion on multiphase flow for completeness. Several review articles on the latter are provided as reference for the interested reader. Among the various pore-scale models, we place the greatest emphasis on pore-network models. Pore-network models are "intermediate-scale", or "mesoscale", models bridging the gap between the pore scale $(1-100 \ \mu m)$ and the core scale (10 cm-1 m). This is an important scale because it incorporates several thousand pores, where most macroscopic behaviors begin to emerge. Network models are also attractive because of their simplicity, which is why they can handle domain sizes much larger than other pore-scale models. However, their simplicity can also be the source of inaccuracy and ambiguity. In some cases, one can modify the simplifying assumptions of traditional network models (at the expense of increasing their complexity) and substantially increase their accuracy, while preserving computational efficiency. In other cases, such a modification may be inherently limited, and a different pore-scale approach has to be sought. We give examples of both scenarios in this chapter. In the second part, we turn our attention to the problem of transferring information obtained from pore-scale models to larger continuum models. This is an essential step, since the latter is the most practical scale of interest to most geochemists, hydrologists, and subsurface engineers. This chapter focuses only on a review of modeling strategies that are designed for scenarios in which the pore and macro scales are tightly coupled, and are not amenable to hierarchical upscaling. Instead these approaches seek to establish a dynamic "two-way communication" between the pore and the continuum scales, typically over a small region of the porous medium. In doing so, both pore-scale and continuum models are used within the same computational domain, which is why they are commonly referred to as "hybrid" models. The hope is that the following material familiarizes beginning researchers with current advances in pore-scale as well as hybrid modeling, and aids them in adopting and improving upon those methods that best fit their research interests.

PORE-SCALE MODELING

Direct pore-scale modeling

Over the past few decades modeling fluid flow and solute transport at the pore scale has seen the adoption and development of various computational methods. The first and perhaps most critical step prior to any modeling effort is the accurate characterization of the pore-space geometry/topology. Imaging techniques such as X-ray microtomography (XMT) (Wildenschild and Sheppard 2013; Noiriel 2015, this volume) have made it possible to obtain accurate 3-D characterizations of the complex pore-space geometry of rock samples. These techniques have also been used to visualize and quantify pore structural changes as a result of biofilm growth (e.g., Iltis et al. 2011) and reactive transport (e.g., Noiriel et al. 2005) as shown in Figure 1. For granular media, Monte Carlo (Maier et al. 2003), cooperative rearrangement (Thane 2006), and sequential sedimentation (Coelho et al. 1997; Øren and Bakke 2002) algorithms are commonly used to digitally reconstruct representations of various grain packs, in which the pore space geometry is well-defined. Such reconstructions, although approximations of actual porous media, provide valuable insights into the link between depositional processes (e.g., cementation and compaction) and hydraulic/transport properties (e.g., permeability) of granular media (Bryant et al. 1993a,b; Bakke and Øren 1997).

Modeling can then proceed either by simulating directly on the complex void geometry, or on a simplification thereof. The former approach is typically referred to as direct modeling, while the latter is closely associated with pore-network modeling (see the *Pore-network models* section). In this section, we review a handful of direct modeling approaches recently used at



Figure 1. (i) 3-D XMT image of biofilm growth on glass beads [Used by permission of John Wiley & Sons, Inc., from Iltis GC, Armstrong RT, Jansik DP, Wood BD, Wildenschild D (2011) Imaging biofilm architecture within porous media using synchrotron-based X-ray computed microtomography. *Water Resources Research*, Vol. 47, no. 2, doi: 10.1029/2010WR009410, Fig. 1. Copyright © 2011 by the American Geophysical Union]. Cross-sectional XMT image of a limestone core sample at (iia) initial state and (iib) difference between initial and final states of a reactive transport experiment (i.e., dissolution by acidic carbon-dioxide-saturated water); (dark) porosity, (high grey level) matrix, (low grey level) dissolved matrix (i.e., increased porosity) [Used by permission of Oil and Gas Science Technology, from Noiriel C, Bernard D, Gouze P, Thibault X (2005) Hydraulic properties and microgeometry evolution accompanying limestone dissolution by acidic water, *Oil & Gas Science and Technology*, Vol. 60, p. 177–192, Fig. 4].

the pore scale. These include computational fluid dynamics (CFD) (Wendt 2009; Molins et al. 2014; Trebotich et al. 2014; Molins 2015, this volume), Lattice-Boltzmann (LB) (Chen and Doolen 1998; Yoon et al. 2015, this volume), and smoothed particle hydrodynamics (SPH) (Monaghan 1992; Tartakovsky et al. 2008b). These methods can be used to solve the governing Equations (1–2) for single-phase, incompressible, Newtonian flow on the pore space.

$$\rho\left(\frac{\partial v}{\partial t} + v \cdot \nabla v\right) = -\nabla p + \mu \nabla^2 v + f_{\rm b} \tag{1a}$$

$$\nabla \cdot v = 0 \tag{1b}$$

$$v|_{\Gamma_{\rm fs}} = 0 \tag{1c}$$

$$\frac{\partial c}{\partial t} + v \cdot \nabla c = D_{\rm m} \nabla^2 c + r_{\rm hm}$$
(2a)

$$-D_{\rm m}\nabla c \mid_{\Gamma_{\rm fs}} = r_{\rm ht} \tag{2b}$$

Equations (1a) and (1b) represent the Navier–Stokes and continuity equations, respectively. v, p, ρ, μ, c denote fluid velocity, pressure, density, viscosity, and solute concentration. f_b denotes body forces exerted on the fluid. Equation (1c) is the no-slip boundary condition at the rock–fluid interface, Γ_{fs} . Since typical Reynolds numbers (Re = $\rho v d/\mu$; d is a characteristic length of the medium typically the grain size) in the subsurface are low (<1), the inertial term (i.e., $v \cdot \nabla v$) is commonly ignored. Equation (2a) represents the species-balance equation with a homogeneous reaction rate of r_{hm} (D_m is the molecular diffusion coefficient), and Equation (2b) represents the heterogeneous reactive boundary condition (of rate r_{ht}) at the rock–fluid interface. While both fluid flow (i.e., Eqn. 1) and solute transport (i.e., Eqn. 2) can be solved via either of the foregoing direct modeling approaches, Eulerian (or mesh-based) methods (e.g., CFD) are typically preferred for computing flow and Lagrangian (or particlebased) methods (e.g., SPH) for transport, since they are devoid of numerical dispersion. Another popular Lagrangian method used for simulating transport at the pore scale is particle tracking (PT). In this method, particles (or random walkers) are propagated through the void space via a deterministic advection step that follows the streamlines, followed by a stochastic diffusion step that obeys Einstein's equation; i.e., $\delta r^2 = 2dD_m \delta t$, where *d* is the dimensionality of the problem, and δr and δt are the stochastic spatial and temporal step sizes, respectively.

Maier et al. (2000, 2003) studied dispersion on digitally generated sphere packs using LB and PT to solve flow and transport, respectively, and reported good agreement with measurements from nuclear magnetic resonance (NMR) spectroscopy experiments. Molins et al. (2012) used an adaptive meshing CFD method to study the "lab-field discrepancy" of geochemical reaction rates (Fig. 2i). Their study provides valuable insight into the vital role of pore-scale modeling as a crucial guide towards predictive macro-scale modeling. Yang et al. (2013) used CFD to solve the Navier–Stokes flow equations on a bead pack and successfully compared the velocity field to that obtained from magnetic resonance velocimetry measurements. Zaretskiy et al. (2010) used a finite-element–finite-volume method to study longitudinal dispersion on a digitized sample of Fontainebleau sandstone. They highlighted the sensitivity of the pore-scale modeling results to the computational mesh employed. Mostaghimi et al. (2012) used a combination of finite difference and PT for flow and transport, respectively, on micro-CT images of Berea sandstone, and produced longitudinal dispersion coefficients in quantitative agreement with experimental data. Ovaysi and Piri (2011) used a Lagrangian approach referred to as the modified moving



Figure 2. (ia) Absolute velocity magnitude, and (ib) total calcium concentration for a two-zone domain of heterogeneously packed calcite grains computed via an adaptive meshing CFD method [Used by permission of John Wiley & Sons, Inc., from Molins S, Trebotich D, Steefel CI, Shen C (2012), An investigation of the effect of pore scale flow on average geochemical reaction rates using direct numerical simulation, *Water Resources Research*, Vol. 48, Issue 3, doi: 10.1029/2011WR011404, Fig. 5]. (iia)-(iic) Distribution of the non-wetting phase in drainage accompanied by thermally-induced melting of the solid phase, for different melting efficiencies described by the Stefan (St) number. LB was used in performing these simulations [Used by permission of Cambridge University Press, from Parmigiani A, Huber C, and Bachmann O, and Chopard B (2011), Pore-scale mass and reactant transport in multiphase porous media flows, *Journal of Fluid Mechanics*, Vol. 686, p. 40-76, Fig. 8]. (iii) Comparison of predicted longitudinal dispersion coefficients (black dots), using the MMPS method, and experiments (red dots) for various Péclet numbers (=ad-vection/diffusion) [Used by permission of Elsevier Ltd., from Ovaysi S, Piri M (2011), Pore-scale modeling of dispersion in disordered porous media, *Journal of Contaminant Hydrology*, Vol. 124, p. 68–81, Fig. 7]

particle semi-implicit (MMPS) method to solve flow and transport on sandstone samples. They obtained longitudinal dispersion coefficients in favorable agreement with experimental data (Fig. 2iii), and included non-inertial effects in their simulations.

Kang et al. (2010) and Yoon et al. (2012) used LB to perform single-phase multi-component reactive transport simulations involving calcite precipitation and dissolution. Kang et al. (2010) additionally pointed out discrepancies between continuum and pore-scale predictions of the reactive front velocity. Multiphase flow and multicomponent reactive transport simulations using direct modeling approaches have also been performed in the literature (Parmigiani et al. 2011; Chen et al. 2013, 2014). For example, Parmigiani et al. (2011) used LB to simulate drainage accompanied by thermally induced melting of the solid phase (Fig. 2ii). Since the focus of this paper is primarily on single-phase transport phenomena, the reader is referred to Meakin and Tartakovsky (2009) for a complete formulation of the governing equations as well as modeling methods for multiphase flow and reactive transport at the pore scale.

Despite the high fidelity of direct modeling predictions and the fundamental insights they provide about the underlying physical mechanisms of a given phenomenon, they often demand high performance computational resources and massive parallelism (Oostrom et al. 2014). This poses a limit on the size and scale of the problems to which they can be applied. Such limitations have made pore-network modeling very popular over the past few decades, which overcome this limitation by simplifying the complex void space geometry while preserving essential features thereof. In the following section, we discuss pore-network models and provide examples of what we mean by "essential features".

Pore-network models

The popularity of pore-network models arose out of the pioneering works of Fatt (1956a,b,c), who studied two-phase drainage on a 2-D lattice network of randomly assigned throat radii. This was a radical shift from the hitherto bundle-of-tubes representation of porous media for computation of various macroscopic properties. Pore networks are simplified representations of the complex pore-space geometry and consist of an interconnected network of pores (or nodes) and throats (or bonds) (Fig. 3i). This geometric simplification is the reason why network models are more computationally efficient than direct pore-scale models, and applicable to much larger domain sizes. The elements of a pore network are typically assigned relatively simple shapes amenable to analytical treatment, e.g., spheres for pores and cylinders for throats. Scenarios in which either pores or throats are assigned zero volumes have also been considered. A summary of various pore/throat shapes used in the literature can be found in Joekar-Niasar (2010). The manner in which pores are connected to their nearest neighbors constitutes the network topology (coordination number is a parameter closely related to the network topology and is defined as the number of connected neighbors to a given pore) while the specific geometric idealizations used to represent pores/throats comprise the network geometry. Proper characterization of both the topological and geometric aspects of porous samples is the first necessary step, if pore-network models are to be predictive.

Early works on pore-network characterization typically involved statistical mappings of pore/throat properties (e.g., radii, coordination number) onto a lattice structure (Payatakes et al. 1980; Mohanty et al. 1987), or adjusting pore/throat properties to match one set of measurements (e.g., capillary pressure) followed by predictions of more difficult-to-measure properties (e.g., relative permeability) (Fischer and Celia 1999; Vogel 2000). However, statistically mapped networks often ignore spatial correlations present in real rocks due to the random assignment of pore/throat properties, and adjusted networks (to match one set of measurements) suffer from non-uniqueness in their representation. For these reasons, statistically generated networks are typically not considered to be predictive. Bryant and coworkers (Bryant and Blunt 1992; Bryant et al. 1993a,b) introduced the concept of physically representative networks, which

marked the beginning of truly predictive network modeling. In their work, permeability, relative permeability, capillary pressure, and permeability–porosity relationships, were all accurately and directly predicted without adjustable parameters using only grain positions in a disordered sphere pack (measured by Finney 1968). The network was extracted via Delaunay Tessellation of the sphere centers (Fig. 3i), and its main attraction was that spatial correlations of the pore space were naturally imbedded into the extracted network. Recent advances in computer and imaging facilities has given rise to various image analysis techniques whereby physically representative networks can be obtained from digitized images of real samples. These include medial-axis (Thovert et al. 1993; Lindquist et al. 1996; Lindquist and Venkatarangan 1999; Prodanovič et al. 2006), watershed (Sheppard et al. 2006; Thompson et al. 2008), and maximal-ball (Silin and Patzek 2006; Dong and Blunt 2009) algorithms. These techniques have resulted in further quantitative predictions of single-phase and multi-phase flow properties in water-wet and mixed-wet media (e.g., Bakke and Øren 1997; Øren et al. 1998; Patzek 2001).

Since Fatt's (1950a,b,c) works, pore-network models have been used in many other applications including: non-Newtonian flow (Lopez et al. 2003; Balhoff 2005), non-Darcy flow (Thauvin and Mohanty 1998; Balhoff and Wheeler 2009), solute dispersion (Sahimi 1986; Sorbie and Clifford 1991; Bijeljic et al. 2004, Bijeljic and Blunt 2007; Acharya et al. 2007b), reactive transport (Hoefner and Fogler 1988; Li et al. 2006; Algive et al. 2010; Kim et al. 2011)



Figure 3. (ia)-(ic) Schematic of network extraction from a granular medium via Delaunay tessellation. The resultant network is essentially a ball-and-stick model, in which the pores and throats are represented here as spheres and cylinders, respectively (ii) Schematic of a two-scale network with half the grains replaced by grain-filling microporosity (due to partial dissolution as is typical in tight gas sandstones) [Figs. 3i and 3ii used by permission of Elsevier Ltd., from Mehmani A, Prodanovič M (2014) The effect of microporosity on transport properties in porous media. Advances in Water Resources, Vol. 63, p. 104–119, Figs. 1 and 3]. (iii) Wormhole formation due to transport-limited matrix dissolution (here a decrease in Damköhler number corresponds to an increase in injection rate) [Used by permission of John Wiley & Sons, Inc., from Hoefner ML, Fogler HS (1988), Pore evolution and channel formation during flow and reaction in porous media, AIChE J, ASME 2011 5th International Conference on Energy Sustainability, Vol. 34, Issue 1, p. 45-54, Fig 12. Copyright © 1988 American Institute of Chemical Engineers]. Simulated supercritical CO₂ saturation distributions for a 50×50 pore-network model of (iva) sandstone and (ivb) shale at $Ca = 10^{-4}$ (i.e., capillary number = viscous forces/capillary forces) [Used by permission of the American Society of Mechanical Engineers, from Ellis JS, Ebrahimi A, Bazylak A (2011), Characterization of a Two-Phase Pore-Scale Network Model of Supercritical Carbon Dioxide Transport Within Brine-Filled Porous Media. In: ASME 2011 5th International Conference on Energy Sustainability, p. 1307-1314, Fig. 7].

(Fig. 3iii), multi-phase flow (Koplik and Lasseter 1985; Al-Gharbi and Blunt 2005; Piri and Blunt 2005; Joekar-Niasar et al. 2010; Ellis et al. 2011) (Fig. 3iv), biofilm growth (Suchomel et al. 1998a,b), and modeling dual-scale media containing micron- and submicron-scale porosity (Mehmani et al. 2013; Mehmani and Prodanovič 2014a,b; Prodanovič et al. 2014) (Fig. 3ii). A larger portion of this literature is devoted to two-phase flow (drainage and imbibition), due to its significance in petroleum engineering (oil recovery) and soil sciences. The reader is referred to comprehensive reviews of pore-scale modeling of multiphase flow by Celia et al. (1995) and Blunt (2001) for more information. An excellent review by Berkowitz and Ewing (1998) further provides insights into the close connection between (quasi-static) two-phase flow and (invasion) percolation theory. It is noteworthy, that while the development of multiphase flow network models as well as single-phase reactive/passive transport network models have separately seen considerable progress over the past few decades, a merger between the two is rarely observed in the literature. Since the geologic storage of CO_2 , for example, involves the flow of supercritical CO_2 and brine in the presence of multicomponent geochemical reactions with the potential for mineral precipitation/dissolution, such a merger is very much needed.

Pore-network models are often regarded as "mesoscale" models acting as a critical bridge between our fundamental understanding of fluid flow/transport physics at the pore scale $(1-100 \ \mu\text{m})$ and our practical desire for modeling at the continuum scale (>1 m). The reliability of such intermediate-scale models, however, hinges upon their underlying simplifying assumptions. There are two levels of approximations made in every pore-network model. The first is geometric, and occurs in replacing the complex void space geometry with an "equivalent" pore network. The other is physical, and occurs when we make simplifying assumptions to describe the pore-scale physics on this idealized geometry. The key to constructing predictive network models is to identify and capture the most essential geometric and physical features associated with a given phenomenon (as alluded to in *Direct pore-scale modeling*). These essential features are *not* invariant from one problem to the next. In multiphase flow, for example, it is very important to include angular geometries in the description of the pore network. Excluding this feature ignores the possibility of corner flow of the wetting phase, and snap-off of the non-wetting phase because of it. On the other hand, including inertial effects in the description of two-phase flow in a pore-network model seems to be of little importance if residual saturation trapped behind an advancing drainage front is to be predicted (Moebius and Or 2014). In single-phase flow, angularities are unimportant and only accurate values for hydraulic conductivities of the network elements are required. When geochemical reactions are present, neglecting to accurately quantify fluid-mineral interfacial areas will lead to large errors in macroscopic predictions. In passive solute transport, while assuming perfect mixing conditions within individual pores is inappropriate for ordered media, it appears to be a good assumption for disordered media (see Network modeling of solute transport).

In identifying these essential geometric and physical features, direct pore-scale modeling as well as controlled (e.g., micromodel) experiments play an indispensable role. Once an essential feature is identified, traditional network models may be modified to attain improved predictive accuracy. Although such a modification may not always be possible, in which case a different mesoscale modeling strategy has to be sought. In the following section, we review network modeling of single-phase solute transport and provide two examples where direct porescale modeling was used to establish the need for modifying traditional mesoscale assumptions. In the first, modifications lead to substantial improvements in the predictive capacity of the traditional network model, while in the second improvements appear to be inherently limited.

Network modeling of solute transport

Pore-network modeling of solute transport has received special interest from many authors in the past few decades and several methodologies have been proposed. A prerequisite to simulating transport is the computation of the velocity field within throats. The procedure is quite standard and involves: a) imposing pressure boundary conditions on the pore network, b) describing flow rates within throats via a constitutive equation e.g., Hagen–Poiseuille for Newtonian fluid in a cylindrical throat, c) enforcing mass balance at each pore e.g., Equation (3), where g_{ij} is the throat hydraulic conductivity, Δp_{ij} the pressure drop across the throat, and N_i^{th} the number of throats connected to pore *i*, d) solving the resultant system of (linear or nonlinear, depending on fluid rheology and/or flow regime) equations for pore pressures, and e) computing throat flow rates/velocities from the aforementioned constitutive throat equation.

$$\sum_{j=1}^{N_i^{\rm th}} \frac{g_{ij}}{\mu} \Delta p_{ij} = 0 \tag{3}$$

The simulation of transport then resumes using computed throat flow rates/velocities. Below in this section, we review several Eulerian and Lagrangian approaches developed in the literature for modeling transport on pore networks. Advantages and limitations of each method are outlined and discussed. We then focus on the main sources of error and difficulty within each modeling class, and the possibility of minimizing them. Finally, we provide a brief review of modeling homogeneous/heterogeneous reactions on pore networks.

Eulerian network models. Bryntesson (2002), Acharya et al. (2005, 2007b), Li et al. (2006), Kim et al. (2011), Mehmani et al. (2012), and Nogues et al. (2013) are among those who have adopted the popular mixed-cell method (MCM), in which solute-balance equations are written for each pore, i.e., Equation (4).

$$V_{p_i} \frac{dc_i}{dt} = \sum_{j=1}^{N_i^{\text{th},q<0}} c_i q_{ij} + \sum_{j=1}^{N_i^{\text{th},q<0}} c_j q_{ij} + \sum_{j=1}^{N_i^{\text{th}}} D_{\text{m}} a_{ij} \frac{\Delta c_{ij}}{l_{ij}} + R(c_i)$$
(4)

In Equation (4), c_i and V_{p_i} are the concentration and volume of pore *i*; q_{ij} , a_{ij} , and l_{ij} are the throat flow rate, cross-sectional area, and length, respectively; $R(c_i)$ is the reaction term. In this method, throats are assumed to be volumeless and the solute within the pores perfectly mixed (i.e., single concentration value assigned to each pore). Solute flow rates within throats are formulated as the algebraic sum of an upwinded advection term (first two terms on the RHS of Eqn. 4) and a linearly varying diffusion term (third term on the RHS of Eqn. 4). Essentially, MCM can be regarded as a low-order, finite-volume method on an unstructured grid, i.e., the pore network. The advantage of MCM is that it is very computationally efficient and highly adaptable to various transport scenarios. For example, Acharya et al. (2005) used MCM to study non-linearly adsorbing solute transport, and determined that more than a million pores were required for their results to be statistically representative. Li et al. (2006) and Kim et al. (2011) studied complex geochemical reaction kinetics of anorthite and kaolinite precipitation/ dissolution relevant to CO_2 sequestration. Nogues et al. (2013) studied porosity/permeability evolutions in carbonates due to carbonic-acid-driven precipitation/dissolution reactions. They considered 18 aqueous species and 5 mineral species undergoing 14 independent reactions. The flexibility and computational efficiency of MCM is why such complex systems acting on sufficiently large pore-scale domains can even be considered.

A number of variants and/or modifications of MCM have also been developed in the literature. For example, Raoof et al. (2013) assign volume to both pores and throats (solute still perfectly mixed in both) and sub-discretize the wetting filaments in the corners of partially drained pores to account for the partial mixing of solute within them (Fig. 4iv). Van Milligen and Bons (2014) proposed a modification to the throat rate expressions used in MCM (i.e., first three terms on the RHS of Eqn. 4) by deriving analytical expressions based on a steady-state plug-flow assumption within throats. They occasionally further sub-discretized throats into smaller "pores" for increased modeling resolution. A generalization of this method



Figure 4. (i) Evolution of a solute plume (three snapshots per case) through a 2-D domain for different Péclet numbers and degrees of heterogeneity simulated using DPT. Fluid injection is from a small region to the left side of the domain marked by a thick segment [Used by permission of John Wiley & Sons, Inc., from Bruderer C, Bernabé Y (2001), Network modeling of dispersion: Transition from Taylor dispersion in homogeneous networks to mechanical dispersion in very heterogeneous ones, Water Resources Research, Vol. 37, p. 897–908, Fig. 6. Copyright © by the American Geophysical Union]. (ii) Comparison of predicted longitudinal (upper curve) and transverse (lower curve) dispersion coefficients using DPT with experiments for unconsolidated granular media, for various Péclet numbers [Used by permission of the Society of Petroleum Engineers, from Bijeljic and Blunt (2006), A Physically-Based Description of Dispersion in Porous Media, In: SPE Annual Tech. Conf. and Exhibition, Society of Petroleum Engineers, SPE-102869-MS, Fig. 3. Copyright 2013, Society of Petroleum Engineers. Further reproduction prohibited without permission]. (iiia) Schematic of partial mixing at the 2-D intersection of two fractures, and (iiib) its conceptualization by a unidirectional flow/transport geometry amenable to analytical treatment [Used by permission of John Wiley & Sons, Inc., modified after Park YJ, Lee KK (1999), Analytical solutions for solute transfer characteristics at continuous fracture junctions, Water Resources Research, Vol. 35, Issue 5, p. 1531–1537, Figs. 1, 2, and 3. Copyright 1999 by the American Geophysical Union]. (iva) Schematic of a drained cubic pore by a non-wetting phase. (ivb) The wetting filaments in the crevices of the pore are subdiscretized into 8 pores (corresponding to the corners) and 12 throats (corresponding to the edges) [Used by permission of Elsevier, Ltd., modified after Raoof A, Nick HM, Hassanizadeh SM, Spiers CJ (2013), PoreFlow: A complex pore-network model for simulation of reactive transport in variably saturated porous media, Computers & Geosciences, Vol. 61, p. 160-174, Fig. 4].

for non-uniform velocity profiles was given by Mehmani (2014) (referred to as the ratemodified MCM), but it was shown that these modifications offered little improvement in model predictions compared to MCM. Algive et al. (2010, 2012) and Varloteaux et al. (2013) similarly modified throat solute rate expressions in the context of reactive transport. These works employed moment theory to derive corrected macroscopic parameters (i.e., solute mean velocity, dispersion coefficient, reaction source term) for each pore/throat element in the long-time asymptotic regime. The model was used to study the effects of dissolution/ precipitation reactions on macroscopic single- and two-phase flow properties in the contexts of CO_2 sequestration and diagenetic alterations in carbonate rocks. Suchomel et al. (1998b) developed a model in which pores were assumed volumeless and throats were sub-discretized into finite-difference grids. Interpore diffusion (equivalent of the third term in the RHS of Eqn. 4) was implicitly incorporated by adjusting numerical diffusion via grid and time-step sizes (although this does not account for diffusion countercurrent to the flow direction), and perfect mixing was assumed within pores. The model was used to study permeability/porosity alterations during bacterial biofilm growth in porous media. A simple but interesting model was developed by Martins et al. (2009), in which the solute balance equations at the pores were formulated as a system of delay-differential equations, i.e., Equation (5).

$$V_{p_i} \frac{dc_i(t)}{dt} = \sum_{j=1}^{N_i^{\text{th},q<0}} c_i(t)q_{ij} + \sum_{j=1}^{N_i^{\text{th},q>0}} c_j(t-\tau_{ij})q_{ij}$$
(5)

In Equation (5), τ_{ij} is the throat residence time, which accounts for the delay in transport from one pore to the next. In essence, the model uses upstream concentrations from earlier time steps to compute solute flow rates flowing into pores at later times. This is implemented by sub-discretizing the throats and marching pore concentrations forward within the subdiscretized segments (akin to a traveling wave). However, several limiting assumptions were made including the neglect of diffusion, plug-flow within throats, and perfect mixing within pores.

A different set of models formulate transport equations in the Laplace transform domain with respect to the time variable. There is a certain extent of elegance and convenience associated with working in the Laplace domain, mainly due to the fact that convolutions of transit-time probabilities are converted to simple algebraic multiplications. Another advantage is that computation of temporal moments becomes rather straightforward (the Laplace transform of transit-time distributions act as moment generating functions). De Arcangelis et al. (1986) developed first-passage-time probabilities for tracer particles that move through throats connecting two neighboring pores. They assumed perfect mixing at the (volumeless) pores and plug flow with molecular diffusion at the throats. Under these conditions, they derived exact transit probabilities in the Laplace domain for particle motions in a network. They then used a "probability propagation" algorithm to determine the first-passage-time distribution of a 10×10 diamond lattice network and computed longitudinal dispersion coefficients for various Péclet numbers (= advection/diffusion). At no point was the timedomain concentration field computed. Koplik et al. (1988) used the same set of equations as de Arcangelis et al. (1986) but diverged in their analysis by writing species balance equations for each pore in the Laplace domain. The resulting linear system of equations was then solved and numerically inverted into the time domain using the Stehfest (1970) algorithm. A strategy for computing higher-order moments of first-passage-time distributions of the network was further outlined. This method was later extended by Alvarado et al. (1997) to reversible adsorption/reaction scenarios, where they arrived at the interesting conclusion that dispersion coefficients depend on the degree of spatial heterogeneity of reactive sites in a porous sample and scale non-linearly with Péclet number. However, they noted that the numerical Laplace inversion step was prohibitive for networks larger than 20×20 pores and inaccurate for large Péclet numbers (>10). Indeed numerical inversion of the Laplace transform is known to be notoriously difficult (often unstable) and ill-posed in computational and applied mathematics. For this reason, although valuable for performing moment analyses, time-domain predictions via these methods on representative sample sizes seem impractical and unlikely.

Lagrangian network models. A more natural description of transport is provided by Lagrangian models, among which particle tracking (PT) is almost exclusively employed. In this method the steady-state flow equation is first solved on the network, as described in the beginning of this section, to obtain mean fluid velocities within each throat. Then, depending on the specific throat geometry, a (rectilinear) velocity profile is assumed and used to track particles from pore to pore, subject to simultaneous convection and molecular diffusion (Fig. 4i). Lagrangian network models are generally more computationally expensive than

Eulerian network models, although more accurate. The limiting factor is the large number of solute particles often required in Lagrangian network models to obtain statistically converged results. PT methods on pore networks can be divided into two categories: a) those that trace particle motions in detail within throats following a discrete-time random walk process (Bruderer and Bernabé 2001; Bijeljic et al. 2004; Acharya et al. 2007a; Jha et al. 2011), and b) those that perform continuous-time random hops from one pore to the next (without explicit throat-level simulations) using throat transit-time distributions (Sahimi et al. 1986; Sorbie and Clifford 1991; Rhodes and Blunt 2006; Bijeljic and Blunt 2006; Picard and Frey 2007). We refer to the first class as DPT (discrete-time particle tracking) and to the second as CPT (continuous-time particle tracking).

Compared to CPT, DPT simulations are more time consuming since computational performance is limited by the time step size (controlled by the minimum throat residence time throughout the network). However, DPT can be quite accurate and has been used to successfully predict dispersion coefficients in unconsolidated granular media (e.g., Bijeljic et al. 2004; Jha et al. 2011) (Fig. 4ii). It is also very flexible in the sense that various velocity profiles (e.g., parabolic or plug-flow) within the throats can be considered, and one has substantial control on how to reassign particles to new outlet throats upon their arrival at the pores. Specifically, particles can be reassigned to any desired throat connected to the arrival pore, and even any desired location on the cross-section of that throat. In contrast, CPT is computationally more efficient but comparatively less flexible and less accurate especially in ordered media, as discussed in *Difficulties and Sources of Error*. The efficiency is due to the fact that particle motions within throats are not explicitly simulated, but are instead imbedded in the throat transit-time distributions. The reduced flexibility and accuracy are due to the loss of control in reassigning incoming particles to accurate cross-sectional locations of the outlet throats of a pore. This means that once an outlet throat is chosen (based on some probability), a transit-time distribution is used to determine the time the particle requires to exit the throat. But the transit-time distribution is typically stationary in time and is derived with uniform inlet conditions for concentration. Therefore, there is no way to specify the cross-sectional location an incoming particle should be assigned to in an outlet throat of a pore. This is particularly important in simulating dispersion in ordered media, for which CPT will not yield the expected (e.g., Edwards et al. 1991) $D_{\rm L} \sim {\rm Pe}_d^2$ scaling, where $D_{\rm L}$ is the longitudinal dispersion coefficient, and Pe_d is the Péclet number defined as the ratio of advection to diffusion (Mehmani 2014).

CPT methods can be further subdivided into those that use deterministic transit-time distributions (e.g., Sorbie and Clifford 1991; Picard and Frey 2007), and those that use ensemble-averaged transit-time distributions (e.g., Bijeljic and Blunt 2006; Rhodes et al. 2009). Bijeljic and Blunt (2006) derived an ensemble-averaged transit-time distribution for Berea sandstone by fitting a truncated power-law distribution to their simulation results. They provided physically meaningful interpretation of the distribution variables and fitted their simulated data with a single adjustable parameter. On the other hand, deterministic transit-time distributions are often derived based on similar mathematics and assumptions as those already discussed in the context of Laplace-domain Eulerian network models (e.g., Rhodes and Blunt 2006; Picard and Frey 2007). These assumptions include perfect mixing within volumeless pores and plug flow within throats. Deterministic transit-time distributions are often derived from the Laplace domain to the time domain in order to draw particle transit times within the throats. An exception to this is Sorbie and Clifford (1991) who derived deterministic transit-time distributions based on rigorous single throat simulations assuming non-uniform velocity profiles.

Difficulties and sources of error. The most common sources of ambiguity and error in both Eulerian and Lagrangian network models of single-phase solute transport are in the

descriptions of partial mixing conditions within pores (Fig. 5a), and shear dispersion (i.e., spreading of the solute due to non-uniform velocity profiles) within throats (Fig. 5b). Accurate modeling of these fundamental transport physics could have a significant impact on quantitative macroscopic predictions of solute dispersion and effective reaction rates. Almost all Eulerian network models developed in the literature seem to assume perfect mixing within pores and neglect shear dispersion within throats. Models assuming Taylor–Aris shear dispersion coefficients within throats are unrealistic because throat lengths are typically not long enough for an asymptotic regime to be reached (Mehmani 2014). On the other hand, the incorporation of shear dispersion into Lagrangian network models (e.g., PT) is quite straightforward (see the discussion of these models above), but accurate description of pore-level partial mixing is still a source of ambiguity in these models.

In reality, perfect mixing occurs when diffusive forces are much more dominant than either advective or reactive forces. In the absence of reactions, increasing the Péclet number changes pore-level mixing conditions from perfect mixing to partial mixing. Figure 5a shows direct numerical simulations of this scenario in a 2-D pore (Mehmani et al. 2014). When reactions are present, increasing the Damköhler number (= reaction/diffusion), while keeping the Péclet number constant, causes a high-concentration-gradient boundary layer to develop near the fluid–solid interface. In pore-network models, the effect of reaction-induced boundary layers may be implicitly taken into account in the calculation of pore concentrations, although this scenario unfortunately has not been explored enough in the literature. In the following, we shall focus on shear dispersion and partial mixing in the absence of chemical reactions, and defer further discussion on the effects and modeling approaches of reactions until the very end of this section.

The ambiguity in accurately describing partial mixing within pores stems from the difficulty in approximating flow streamlines within pores and the extent of diffusive mixing occurring therein (Fig. 5a). Sahimi et al. (1986) and Bruderer and Bernabé (2001) developed simple intuitive rules for the redistribution of solute particles from the inlet to the outlet throats of a 2-D cross-shaped volumeless pore. Both ignored diffusive mixing within pores, in the sense that the particles could not randomly "hop" between outlet throats. Sorbie and Clifford (1991) introduced general heuristic, and admittedly approximate, rules for redistributing particles arriving at a pore among its outlet throats. These rules, and variants thereof, were subsequently



Figure 5. (a) Schematic of partial mixing within a sample pore and its dependence on local Péclet number (underlying streamlines also shown) [Used by permission of John Wiley & Sons, Inc., modified after Mehmani Y, Oostrom M, Balhoff MT (2014), A streamline splitting pore-network approach for computationally inexpensive and accurate simulation of transport in porous media, *Water Resources Research,* Vol. 50, Issue 3, pp. 2488–2517, Fig. 1. ©2014. American Geophysical Union. All Rights Reserved]. (b) Schematic of shear dispersion of solute within the throats of a 1D pore-network, due to the non-uniform velocity profile (shown by curved solid line). Figures obtained through CFD simulations on respective pore-space geometries using COMSOL© [Used by permission of author, modified after Mehmani (2014), Modeling of single-phase flow and solute transport across scales, PhD dissertation, University of Texas at Austin, Fig. 4.5].

applied and analyzed in later publications (e.g., Acharya et al. 2004, 2007a,b; Bijeljic et al. 2004; Bijeljic and Blunt 2007). At high Péclet numbers, these rules reduce to redistribution of solute particles based on flow-rate-averaged probabilities; at low Péclet numbers, they reduce to redistribution based on throat cross-sectional-area-averaged (multiplied by other corrective parameters) probabilities. These rules, however, are quite inaccurate at moderate-to-high Péclet regimes as they completely ignore the underlying streamlines that map the inlets to the outlets of a given pore (Mehmani et al. 2014). Jha et al. (2011) proposed a deterministic mapping of incoming particles to the outlet throats, intendedly along the streamlines, in an attempt to generalize the rules developed by Bruderer and Bernabé (2001). The approach was limited to pores with a coordination number of four or less, and ignored mixing due to molecular diffusion. Despite being attractive due to its simplicity, the mapping neglected spatial orientations of the inlet throats, and consequently the interaction between inflowing streams, and was therefore less predictive (Mehmani et al. 2014).

The mixing problem within pores has additionally received quite a bit of attention from the field of fracture-network modeling (e.g., Berkowitz et al. 1994, Park and Lee 1999; Park et al. 2001a,b; Johnson et al. 2006). Although in this context, the "pores" correspond to the juncture at which two fractures intersect. Park and Lee (1999) derived physically sound, efficient, and accurate transition probabilities for mapping particles from the inlets to the outlets of fracture junctions. In doing so, they conceptualized partial mixing at fracture junctions as a one-dimensional problem amenable to analytical treatment (Fig. 4iii). While accurate for fracture networks, the mixing equations of Park and Lee (1999) are not applicable to 3-D networks of porous media. This is because fracture junctions are essentially 2-D cross-shaped "pores", which are geometrically simpler than the 3-D pores with non-planar throat orientations found in porous media.

In an effort to accurately capture partial mixing within pores and shear dispersion within throats, Mehmani et al. (2014) and Mehmani (2014) developed two pore-network models, respectively: the streamline splitting method (SSM) and superposing transport method (STM). These models were developed with the aim of exploring the possibility of capturing the porescale physics discussed above under an Eulerian framework. The insistence on the models being Eulerian was because, in the context of pore-network modeling, such a framework is generally computationally more efficient than a Lagrangian counterpart, albeit possibly less accurate. Therefore, the goal was to find an appropriate balance between acceptable predictive accuracy and computational efficiency. To satisfy the latter, both models used MCM as the starting point for their development. In MCM, a single concentration value is assigned to every pore, which implicitly assumes perfect mixing within them, independent of the local Péclet regime. Speciesbalance equations are then written in every pore, and solved for the pore concentrations. To circumvent the perfect mixing assumption, SSM divides the volume of the pores into smaller compartments (or pockets) and assigns different concentration values to each compartment. The number of compartments within a pore is equal to the number of inlets of that pore, as each inlet may carry with it fluids with different concentrations that need to be kept separate. However, mass transfer (or mixing) between two adjacent compartments within the same pore is allowed, which takes place purely due to molecular diffusion (perpendicular to the streamlines). Unlike Park and Lee (1999), pore-wall effects on the mixing process were also taken into account. The outlets of each compartment are determined by "splitting" the inflowing "streams" between the outlet throats, using a constrained optimization algorithm; which was shown to be also applicable in Lagrangian network models (Mehmani et al. 2014). The compartments are then regarded as control volumes and species balance is enforced in all of them. Figure 6 provides a conceptual picture underlying SSM and contrasts it to that of MCM. It was shown that SSM predictions were in very good agreement with direct CFD simulations as well as micromodel experiments, and computational costs were only marginally higher than MCM.



Figure 6. Schematic of a single pore p0 connected to four neighbors (whose own neighbors are not depicted). (a) Conceptual picture of MCM, where pores are considered as perfectly mixed reactors (to which single concentration values are assigned). (b) Conceptual picture of SSM, where pore volumes are divided into smaller compartments (or *pockets* denoted by pk) in order to account for partial mixing within pores (the depiction is an abstract, rather than literal, representation of the splitting of streamlines) [Used by permission of John Wiley & Sons, Inc., modified after Mehmani Y, Oostrom M, Balhoff MT (2014), A streamline splitting pore-network approach for computationally inexpensive and accurate simulation of transport in porous media, *Water Resources Research*, Vol. 50, Issue 3, p. 2488–2517, Fig. 2. ©2014. American Geophysical Union. All Rights Reserved].

On the other hand, the focus in STM is placed on shear dispersion due to the non-uniform velocity profiles within throats. This important pore-scale process is ignored in MCM as it assumes throats to be volumeless, which causes the solute to instantaneously transport from one pore to the next. In STM, elementary species rate expressions are constructed for each throat, as a function of their aspect ratio (i.e., diameter over length). These elementary rate expressions are then used to perform space-time superpositions across the pore network. This is accomplished by recording pore concentrations as they dynamically evolve during a simulation, which makes it similar to, but much more general than, the model developed by Martins et al. (2009) (the latter ignores shear dispersion and molecular diffusion). The superpositions in STM have the effect of performing convolutions of the elementary rate expressions across the pore network. This makes STM useful in other fields such as signal transmission in electrical engineering. The obvious limitation of STM is that it is applicable to linear transport problems only. This excludes scenarios with non-linear fluid-mineral reactions such as those encountered in CO_2 sequestration. But, for linear problems, the method was verified against direct CFD simulations and simple convolution integral expressions. It was additionally validated against longitudinal dispersion experiments for unconsolidated bead packs from the literature.

Yet a final question remains to be answered: what is the actual magnitude of the improvements imparted to the prediction of a parameter of interest by incorporating these additional pore-scale details into our network models? After all, modeling efforts should ideally seek ultimate simplicity while disposing of unnecessary complexity. Mehmani et al. (2014) showed that the impact of pore-level mixing assumptions, partial vs. perfect, is very significant in ordered media (e.g., micromodels), but comparatively insubstantial in 3-D disordered granular media (e.g., sandstones). In fact, the average difference between the concentration fields obtained via SSM and the relatively less accurate MCM for disordered granular media was ~6% of the maximum concentration value (although the impact on upscaled transverse

dispersion coefficients is yet to be studied). A similar conclusion was drawn by Park et al. (2001a,b), who used PT to study the impact of various mixing assumptions (perfect mixing vs. no mixing) at fracture junctions on overall macroscopic transport behavior in ordered and disordered (2-D) fracture networks. They found that mixing assumptions have a larger impact in ordered media compared to random media (i.e., less than 5% of fracture junctions were affected in random media), and they attributed this to the lower effective coordination number and higher inlet flux ratios, at fracture junctions, in random networks. Despite the various differences between the two media (porous vs. fractured), the conclusions appear to be the same: for disordered (granular/fractured) media the model with the simpler description of pore-level mixing is the one that is preferred (e.g., MCM).

On the other hand, Mehmani (2014) showed that the incorporation of shear dispersion within throats in network models of disordered granular media is rather important at high Péclet numbers (>100). Namely, longitudinal dispersion coefficients predicted by STM undergo a ~2.5-fold improvement compared to MCM. This result corroborates earlier works that used PT to simulate longitudinal dispersion (Acharya et al. 2007a; Jha et al. 2011). A more important conclusion drawn using STM was that all Eulerian network models, including STM, are inherently limited at sufficiently high Péclet numbers when applied to ordered media (e.g., micromodels). This was demonstrated by simply conceptualizing a straight tube under purely advective transport, as an equivalent 1D "network" (or string) of shorter tubes (Fig. 7a). The concentration profiles obtained from STM and MCM were shown to be representative of neither of the analytically known flux-averaged or cross-sectional-averaged concentration profiles (Fig. 7b). Even worse was the fact that simulated profiles were shown to converge towards a Gaussian distribution with an increase in the number of tube segments traveled. This is known to be impossible under a purely advective regime (Fig. 7c). It is known from Taylor-Aris theory that in the presence of molecular diffusion (no matter how small its contribution) a $D_{\rm L} \sim {\rm Pe}_d^2$ asymptotic scaling is expected from this system ($D_{\rm L}$ is the longitudinal dispersion coefficient, and Pe_d is the Péclet number); similar to ordered media (Edwards et al. 1991). Instead, an Eulerian network model (e.g., STM) would yield a scaling that is closer to $D_{\rm L} \sim {\rm Pe}_d$ at high ${\rm Pe}_d$, since Gaussian profiles are reached not because of molecular diffusion, but because of cross-sectional smearing of concentrations at the "pores"



Figure 7. (a) Schematic of a circular tube under pure advection (constant inlet concentration), divided into segments along the dashed lines. (b) Concentration profiles (at different times) along the duct, including: true cross-sectional average (dashed-dotted line), true flux-averaged (dashed-plus line), simulated STM (thick solid line), and simulated MCM (thin solid line). Arrow of time annotated, and λ is *x*-normalized against the length of 50 segments. (c) Convergence of STM profiles towards a normal distribution, with an increase in the number of tube segments traveled (annotated by N_J). *x* is normalized by the distance travelled by the (tube) centerline velocity (= V_0t) [Used by permission of author, from Mehmani (2014), Modeling of single-phase flow and solute transport across scales, PhD dissertation, University of Texas at Austin, Fig. 4.12].

(denoted by dashed lines in Fig. 7a). In other words, solute particle "memories" are effectively erased upon their arrival at the pores (i.e., Markovian process), whereas in reality they are retained over a given distance that depends on the local Péclet regime. This distance is larger than the length of a single throat at sufficiently high Péclet numbers. The foregoing problem is not just specific to Eulerian network models. Lagrangian methods that draw throat transit times from the *same* distribution, that is stationary in time, are equally limited. An example includes the CPT model developed by Sorbie and Clifford (1991). The implication is that the "simplest" model, capable of accurately capturing shear dispersion in ordered media, must preserve particle memories in passing them from the inlet to the outlet throats of a pore. This means that it is not just sufficient to know how long a particle takes to transit a given throat, but also the cross-sectional position of that throat whence it exits. Such detailed information seems only resolvable in a pore network, with adequate simplicity, under a Lagrangian framework.

Reactions. Comprehensive reviews on modeling reactive transport can be found in Steefel and MacQuarrie (1996), Steefel et al. (2005, 2013). Reactions at the pore-scale can be divided into those occurring within the fluid bulk, referred to as homogeneous reactions, and those that take place at the fluid-mineral interface, referred to as heterogeneous reactions. Homogeneous reactions themselves consist of zero-order decay processes independent of species concentrations, and higher-order concentration-dependent reactions that occur due to the mixing of two or more solute species. Homogeneous reactions are commonly modeled as instantaneous processes at equilibrium, while heterogeneous reactions as kinetically controlled processes. In direct pore-scale models, the former is described via source terms in the solute balance equation (Eqn. 2a), and the latter via appropriate boundary conditions at the fluid-mineral interface (Eqn. 2b). In Eulerian network models, where pores are viewed as continuously stirred tank reactors, all reaction types (as well as adsorption) are almost invariably described via source terms in the mesoscale solute balance equation (Acharya et al. 2005, 2007b; Li et al. 2006, 2007a,b; Kim et al. 2011, 2012; Mehmani et al. 2012; Raoof et al. 2013; Nogues et al. 2013). This effectively ignores concentration gradients and transportlimited effects on reaction rates at the scale of individual pores. Depending on the application, this may or may not be a good assumption. Li et al. (2008) showed that for typical flow and chemical/mineralogical conditions relevant to geologic CO₂ sequestration, pore-level perfect mixing appears, for all practical purposes, appropriate. Their conclusion was based on molecular diffusion coefficient and activity of the hydrogen ion among other assumptions. In problems where these parameters assume very different values, transport-limited effects can become important and the perfect mixing assumption inappropriate.

The network models developed by Algive et al. (2010, 2012) and Varloteaux et al. (2013) are the only known exception, in which concentration gradients due to fluid-mineral reactions are implicitly taken into account by calculating effective transport parameters for the pores and throats. However, their model is limited to long-time asymptotic regimes, where changes in the concentration field are slow with time. Moreover, none of the above mentioned models accounts for diffusion-limited mixing reactions that occur when two fluids with differing chemical compositions come into contact within a pore. In SSM, discussed in Difficulties and sources of error, the rate of mass transfer between two adjacent compartments within a pore (*pk1* and *pk2* in Fig. 6b) is computed via solving a local bounded Riemann problem. Mehmani et al. (2014) proposed the possibility of modifying this local problem (via appropriate initial/boundary conditions) to derive mass transfer rates in the presence of mixing-induced reactions. For reactions at the fluid-solid interface within the throats, one may derive modified elementary rate expressions that can be used in the STM model discussed in Difficulties and sources of error (although limitations for ordered media still persist). This would complement the long-time asymptotic method of Algive and coworkers, for short-time pre-asymptotic regimes. Finally, particle tracking (PT) can be used to describe all foregoing

reactions types at the pore scale in high detail. This requires the development of non-trivial stochastic particle interaction criteria and collision probabilities between fluid–fluid and solid–fluid species (Gillespie 1977; Benson and Meerschaert 2008; Dentz et al. 2011). However, the adoption and implementation of PT in pore networks for accurate simulations of geochemical reactions have unfortunately been scarce.

HYBRID MODELING

Flow and reactive transport phenomena in the subsurface often span a wide range of spatial scales (nanometer to kilometer), which renders the development of predictive models capable of accurately bridging them a formidable task. Macroscopic parameters (e.g., permeability) or closure relations (e.g., capillary pressure vs. saturation) are commonly extracted from pore-scale modeling or experiments on representative samples of the real medium, followed by their direct implementation into continuum-scale simulators. While such an approach is appropriate for situations in which a clear separation between scales exists, it may lead to large errors otherwise (Kechagia et al. 2002; Tartakovsky et al. 2008b; Battiato and Tartakovsky 2011). Mixing induced precipitation ensuing form a bimolecular reaction is one such example (Battiato et al. 2009). In these cases a *dynamic* communication between the pore and the continuum scales seems to be required.

Li et al. (2006, 2007a,b) used a pore-network model to study reaction kinetics of kaolinite and anorthite in the context of geologic carbon sequestration. They demonstrated that reaction rates obtained from continuum-scale representations of transport and/or using volume/fluxaveraged concentrations in reaction rate expressions leads to large errors, sometimes even wrongly predicting the direction of the reactions (i.e., precipitation vs. dissolution). Although their study was qualitative in nature, using 3-D regular lattice networks, it provided an explanation for the commonly reported discrepancy between reaction rates observed at the field scale and those obtained from well-mixed batch experiments on crushed samples. Namely, transport limitations at the pore scale control overall reaction rates, which are non-existent under well-mixed conditions in batch experiments. Effects of flow rate and reactive cluster size/ abundance were also studied, and it was concluded that the higher the degree of incomplete mixing (i.e., spatial variability of concentration) the higher the scaling error. Incomplete mixing was found to be strongest at medium flow rates (or Péclet numbers). Kim et al. (2011) and Kim and Lindquist (2012) extended the work of Li and coworkers using networks extracted from XMT images of real sandstones. They were able to determine surface mineral distributions from the images allowing for better quantitative analysis. Similar conclusions were drawn regarding the "lab-field discrepancy", and an approximately power-law scaling of reaction rates vs. flow rate was reported for anorthite, while a more complex scaling emerged for kaolinite. In the context of filtration combustion in porous media, Lu and Yortsos (2005) similarly observed that spatially averaged macroscopic reaction rates were very different (discrepancies of a factor of two or higher) than those determined from using averaged variables in microscopic reaction rate expressions. They attributed this to the strong influence of microscopic heterogeneities on macro scale behavior. Recently, Molins et al. (2012) conducted sophisticated direct pore-scale simulations of calcite dissolution, and showed that pore-scale heterogeneities can result in an underestimation of reaction rates, due to mass transport limitations, even when total reactive surface area and porosity are held constant between samples.

Kechagia et al. (2002) demonstrated that for reactive transport with fast/finite kinetics, homogenization of microscopic equations via volume averaging does not hold except at the limit of macroscopic equilibrium. They further showed that even under these circumstances an eigenvalue problem enforces a coupling between the micro and the macro scales. A systematic study was performed by Battiato and Tartakovsky (2011) to identify transport regimes

(characterized by the Damköhler (Da) and Péclet (Pe) numbers), in which a purely continuum description of advection-diffusion with nonlinear surface reactions breaks down. They used multiple-scale expansions to upscale the pore-scale equations, and presented their results in the form of a Pe–Da phase diagram. The results were in agreement with an earlier work that used volume averaging (Battiato et al. 2009), which substantiated their independence from the specific upscaling method employed. A recent work by Boso and Battiato (2013) extended this analysis to three-component systems undergoing two homogeneous and one heterogeneous reactions (all reversible). It is interesting to note, that Molins et al. (2012) located their simulations on the aforementioned Pe–Da phase diagram and determined their correspondence to a case in which spatial scales were coupled. While all foregoing examples focused on geochemical reactions, a coupling between scales is also observed in problems involving viscous and density-driven instabilities in multiphase and miscible displacement among others (Tartakovsky et al. 2008c).

The implications of the foregoing studies (among others) have given rise to a new class of modeling approaches referred to as "hybrid multiscale methods" (Scheibe et al. 2007), in which micro- and macro-scale simulations are simultaneously performed on the same computational domain. Balhoff et al. (2007) were one of the first to couple a pore-scale model to the continuum. However, their iterative coupling strategy had strong limitations in terms of flexibility and efficiency. These limitations were later lifted by Balhoff et al. (2008) through the introduction of *mortars*. Mortars are finite-element function spaces that ensure the (weak) continuity of flux at the interface between two coupled models (e.g., pore-scale and continuum-scale), and are the essential component of a highly flexible, efficient, and accurate non-overlapping domain decomposition method (Bernardi et al. 1994; Arbogast et al. 2000; Peszynska et al. 2002). Subsequently, Sun et al. (2012a) showed that mortars can be used as accurate upscaling tools for pore-scale models in obtaining macroscopic properties (e.g., permeability). They demonstrated that a large heterogeneous pore-scale domain can be decomposed along structural discontinuities and coupled via mortars to closely approximate the true permeability. Sun et al. (2012b) developed a single-phase reservoir simulator, in which Darcy grids in the near-well region were substituted with pore-scale models (Fig. 8ii). The study focused on upscaling strategies for the permeability field of the near-well region. In the foregoing studies, application of mortars was limited to linear, single-phase Newtonian flow without species transport and computational aspects were left unexamined. These issues were later addressed by Mehmani et al. (2012) and Mehmani and Balhoff (2014), who extended the use of mortars to non-linear (power-law) flow and (passive/reactive) solute transport and demonstrated computational efficiency and parallel scalability of their new algorithms (Fig. 8iii). The application of the above hybrid mortar methods is most appropriate when a tight coupling between spatial/temporal scales exists and a characterization of the pore-scale subdomain is available. This scenario is most likely to occur in a "skin-deep" region around wellbores (e.g., matrix acidization, and CO₂ leakage through wellbore cement), in which fluids are furthest away from equilibrium and direct access to porescale samples for geometric characterization is available.

Tartakovsky et al. (2008a) non-iteratively coupled the pore scale and the continuum in a diffusion-reaction system using an SPH formulation for both scales. They were able to show that their hybrid method produced reliable predictions, compared to modeling purely at the pore scale, of mixing-induced precipitation, which is typically not amenable to a purely continuum description as the reactions occur over the length of a few grain diameters (Fig. 8i). While the non-iterative nature of the method makes it very attractive from a computational standpoint, the method lacks flexibility in the sense that it requires both scales to be formulated using SPH. Advection was additionally ignored in that work. A novel overlapping method for coupling porescale inclusions to the surrounding continua was developed by Battiato et al. (2011) and was successfully verified for the case of Taylor dispersion in a reactive fracture. While the method



Figure 8. (i) Mixing-induced mineral precipitation (represented by black particles) obtained from SPH simulations on a (ia) fully pore-scale domain, and (ib) a hybrid domain with a pore-scale region sandwiched between two continuum regions [Used by permission of the Society for Industrial and Applied Mathematics, from Tartakovsky AM, Tartakovsky DM, Scheibe TD, Meakin P (2008), Hybrid simulations of reaction-diffusion systems in porous media, SIAM Journal on Scientific Computing, Vol. 30, p. 2799–2816, Fig. 5. Copyright ©2008 Society for Industrial and Applied Mathematics. Reprinted with permission. All rights reserved]. (iia) Pressure, and (iib) flow fields obtained from the coupling of a nearwell pore-scale region to the surrounding continuum using mortars [Used by permission of the American Chemical Society, from Sun et al. (2012b), Hybrid Multiscale Modeling through Direct Substitution of Pore-Scale Models into Near-Well Reservoir Simulators, Energy & Fuels, Vol. 26, p. 5828–5836, Fig. 6 and 7. Copyright © 2012, American Chemical Society]. (iiia) Pressure, and (iiib) concentration fields obtained on a (3-D) hybrid domain (results collapsed onto 2-D) with non-matching continuum grids [Used by permission of the Society for Industrial and Applied Mathematics, from Mehmani Y, Balhoff MT (2014), Bridging from pore to continuum: A hybrid mortar domain decomposition framework for subsurface flow and transport, Multiscale Modeling & Simulation, Vol. 12, p. 667-693, Fig. 11. Copyright © 2014 Society for Industrial and Applied Mathematics. Reprinted with permission. All rights reserved]. (iv) Schematic of the heterogeneous multiscale method (HMM) for performing hybrid two-phase drainage simulations [Used by permission of Springer and Advances in Applied Mathematics, Modeling, and Computational Sciences, from Chu J, Engquist B, Prodanovič M, Tsai R (2013), A multiscale method coupling network and continuum models in porous media II-Single-and two-phase flows. In: Advances in Applied Mathematics, Modeling, and Computational Sciences, p. 161–185, Fig. 4. With kind permission from Springer Science and Business Media]. Shaded rectangles represent network models, which are used to compute the pressure field (those depicted at the boundaries of macroscopic control volumes) and advance the saturation front (those oriented normal to the displacement front).

is more general compared to the previous work (it included advection and was not limited to SPH), it appears to be limited to rather small pore-scale point inclusions that are dependent on the underlying macro-grid structure of the domain. The specific coupling algorithm used can also be rather computationally expensive, since the differential-algebraic system was solved by iterating between the differential and algebraic parts, rather than solving the system as a whole. Roubinet and Tartakovsky (2013) built upon the work of Battiato et al. (2011) and developed a non-overlapping hybrid method in 1D, in which the computational performance was enhanced by solving a global (instead of a sequential) differential-algebraic system. The method used finite volumes to discretize both the continuum- and the pore-scale subdomains, and introduced

additional unknowns for each grid on either side of the interface. This dependence of interface unknowns on the number of interface grids can render the approach quite expensive for large problems. It is also noteworthy, that the approach seems to be a special case of the Global Jacobian methods of Ganis et al. (2012) and Mehmani and Balhoff (2014).

Chu et al. (2012) proposed an approach based on the heterogeneous multiscale method (HMM), in which macroscopic conservation equations are written assuming unavailability of constitutive flow equations (e.g., Darcy's law) at the macro scale. The missing data were instead supplied from locally sampled pore-network simulations across the domain. Specifically, pore-scale models were used as providers of accurate in/out-fluxes at the boundaries of macroscopic control volumes. Chu et al. (2013) extended this work to twophase flow problems, where pore-network models were used to track macroscopic two-phase drainage fronts (Fig. 8iv). While the method is a superior alternative to pure continuum-scale models in cases where local effects are dominant, its extension to problems with strong fluidmineral reactions (where a macroscopic source/sink term is unknown and highly coupled to transport itself) is not obvious. Sheng and Thompson (2013) developed a dynamic coupling algorithm for two-phase flow, in which network models were embedded at the center of 1D macroscopic control volumes. The network models provided relative permeability values to the continuum model, while the continuum model provided fractional flow rates as boundary conditions to the network model. The network model was specially designed to handle these non-trivial boundary conditions. The large time-scale discrepancy between the pore scale and the continuum scale was addressed by performing successive steady-state simulations at the pore scale. The work showed that macroscopic and pore-scale saturations do not necessarily agree, and highlighted the difficulty in enforcing such an agreement. Tomin and Lunati (2013) developed a hybrid algorithm based on the multiscale finite volume (MsFV) method. In this approach the macroscopic domain is discretized into primal control volumes, upon which a complementary dual grid is superposed. Local pore-scale problems are then solved on the dual grid, in order to construct a set of bases as well as correction functions; these can be thought of as pore-scale closures for the continuum model. Global balance equations are then solved by forming a linear combination of these bases/correction functions and computing their corresponding multipliers. The continuum solution is then used to construct a conservative velocity field at the pore scale for advancing tracer concentrations and/or phase saturations. The hybrid method was successfully verified against fully pore-scale simulations for tracer transport and stable/unstable drainage problems.

The essence of all hybrid methods discussed is a "two-way communication" between the pore scale and the continuum. Hybrid methods are a relatively recent development compared to single-scale modeling strategies such as molecular dynamics (MD), pore-scale modeling, and reservoir simulation. Naturally, their development has been based upon the vast diversity of single-scale methods and various combinations thereof. Consequently, they themselves are very diverse in the ways they approach the problem, which might raise the obvious question as to which hybrid method is most appropriate for a given problem. Scheibe et al. (2015) have attempted to provide a general road map for choosing an appropriate hybrid strategy, referred to as the Multiscale Analysis Platform (MAP), depending on the degree of complexity of the hydrological problem at hand (i.e., degree of separability between temporal/spatial scales). MAP classifies various hybrid methods into separate "motifs", and leads its user (with a specific application in mind) towards a suitable motif by asking a series of questions regarding the nature of the problem to be solved (Fig. 9). Such a classification additionally provides a useful context, within which the ever-increasing multiscale methods developed in the literature can be categorized.



Figure 9. Flowchart of the multiscale analysis platform (MAP) [Used by permission of John Wiley and Sons, from Scheibe TD, Murphy EM, Chen X, Rice AK, Carroll KC, Palmer BJ, Tartakovsky AM, Battiato I, Wood BD (2015), An analysis platform for multiscale hydrogeologic modeling with emphasis on hybrid multiscale methods, *Groundwater*, Vol. 53, p. 38–56, Fig. 3. © National Ground Water Association].

CONCLUSIONS

Recent advances in rock imaging technology (e.g., XMT) have made it possible to resolve the complex pore-space geometry in sufficient detail, to the point that direct numerical simulations of various pore-scale processes are becoming commonplace and an area of active research. In addition, novel micromodel fabrication and measurement techniques (Werth et al. 2010; Karadimitriou and Hassanizadeh 2012) are allowing for an unprecedented possibility of quantitative comparison between modeling and experiments. However, our recent capability to resolve much of the pore-scale details with high accuracy, both experimentally and numerically, does not detract from the importance of mesoscale (or pore-network) models and the role they play in bridging the pore scale to the continuum. On the contrary, we can now use observations from such detailed experimental and direct numerical results to increase the accuracy of traditional network models and remove some of their ambiguity. The significance for doing so is threefold: a) it allows us to translate our detailed observations to an intermediate scale that is simple enough to understand, b) forces us to determine what features of the problem are the most essential so we can discard unnecessary details, and c) regardless of advances in computational performance, mesoscale models will always be one scale ahead of direct porescale methods. While improving network models based on detailed numerical/experimental observations is often successful, there are however limitations to this process. In the first part of the chapter, we tried to emphasize this point by providing two examples in the context of single-phase solute transport.

In the first example, we demonstrate that modifications to a traditional Eulerian network model allows it to capture partial mixing within pores with higher accuracy, while preserving computational efficiency and relative simplicity. In the same example, we also show that the need for accurately capturing partial mixing within pores depends on the porous medium. If the porous medium is ordered, partial mixing is important. But in disordered media (e.g., sandstones), partial mixing does not seem to be as significant, which means that we do not need to add this additional complexity to our mesoscale model. In the second example, we similarly show that modifications of a traditional Eulerian network allows it to capture shear dispersion within throats, and improves predictions of macroscopic dispersion in disordered (granular) media. However, when the porous medium is ordered, no amount of modification appears to prevent it from producing the wrong D_L vs. Pe_d scaling due to the Eulerian nature of the network model. Thus, a Lagrangian network model seems to be the next simplest mesoscale description. Similar conclusions may or may not hold for problems involving multiphase flow, geochemical reactions, adsorption, etc. However, the current state of computational power and experimental advances are allowing for the possibility of investigation into these questions more than ever.

On the other hand, the practical need to perform large-scale predictions requires pore-scale information to be translated to the continuum scale. While traditional upscaling approaches are successful in various scenarios, their applicability is rather limited when insufficient separation between scales exists. Overwhelming evidence in the literature hints towards this being quite common in the presence of strong fluid-mineral reactions; rendering averaged reaction rates transport-limited at the pore scale. Another example involves viscous and density-driven instabilities in multiphase and miscible displacement scenarios (Tartakovsky et al. 2008c). In recent years, various "hybrid" models have been developed as a means of simulating such scenarios, with the typical premise that the break-down of macroscopic continuum equations occurs locally. The essence of all such methods is a "two-way communication" between the pore and the macro scales, accomplished by incorporating models of both scales into the same computational domain. A useful classification given by Scheibe et al. (2015) facilitates one in choosing a suitable hybrid method, from a large selection developed in the literature, depending on the particular application at hand. As developments in both pore-scale and hybrid models continue into the future, the hope is for all these efforts to converge towards a deeper understanding of the scale transition problem and arrive at a predictive description (modeling and/or theory) of the relevant processes. Such an understanding would undoubtedly be indispensable in tackling currently pressing and highly challenging problems such as the safe sequestration and storage of anthropogenic CO₂, increasing hydrocarbon recovery from mature formations, and clean-up and remediation of nuclear waste repositories.

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REFERENCES

Acharya RC, Van Dijke MIJ, Leijnse A, Van der Zee SEATM, Sorbie KS (2004) Upscaling of tracer transport including convection and Brownian motion using a 3D network model. Dev Water Sci 55:115–125.

Acharya RC, Van der Zee SEATM, Leijnse A (2005) Transport modeling of nonlinearly adsorbing solutes in physically heterogeneous pore networks. Water Resour Res 41:W02020, doi: 10.1029/2004WR003500

- Acharya RC, Van Dijke MIJ, Sorbie KS, Van der Zee SEATM, Leijnse A (2007a) Quantification of longitudinal dispersion by upscaling Brownian motion of tracer displacement in a 3D pore-scale network model. Adv Water Resour 30:199–213
- Acharya RC, Van der Zee SEATM, Leijnse A (2007b) Approaches for modeling longitudinal dispersion in pore-networks. Adv Water Resour 30:261–272

Al-Gharbi MS, Blunt MJ (2005) Dynamic network modeling of two-phase drainage in porous media. Phys Rev E 71:016308

- Algive L, Bekri S, Vizika O (2010) Pore-network modeling dedicated to the determination of the petrophysical-property changes in the presence of reactive fluid. SPE J 15:618–633
- Algive L, Békri S, Nader FH, Lerat O, Vizika O (2012) Impact of diagenetic alterations on the petrophysical and multiphase flow properties of carbonate rocks using a reactive pore network modeling approach. Oil Gas Sci Technol 67:147–160
- Alvarado V, Davis HT, Scriven LE (1997) Effects of pore-level reaction on dispersion in porous media. Chem Eng Sci 52:2865–2881
- Arbogast T, Cowsar LC, Wheeler MF, Yotov I (2000) Mixed finite element methods on nonmatching multiblock grids. SIAM J Numer Anal 37:1295–1315
- Bakke S, Øren PE (1997) 3-D pore-scale modelling of sandstones and flow simulations in the pore networks. SPE J 2:136–149
- Balhoff MT (2005) Modeling the flow of non-Newtonian fluids in packed beds at the pore scale. PhD dissertation, Louisiana State University
- Battiato I, Tartakovsky DM (2011) Applicability regimes for macroscopic models of reactive transport in porous media. J Contam Hydrol 120:18–26
- Balhoff MT, Wheeler MF (2009) A predictive pore-scale model for non-Darcy flow in porous media. SPE J 14:579–587
- Balhoff MT, Thompson KE, Hjortsø M (2007) Coupling pore-scale networks to continuum-scale models of porous media. Comput Geosci 33:393–410
- Balhoff MT, Thomas SG, Wheeler MF (2008) Mortar coupling and upscaling of pore-scale models. Computat Geosci 12:15–27
- Battiato I, Tartakovsky DM, Tartakovsky AM, Scheibe T (2009) On breakdown of macroscopic models of mixing-controlled heterogeneous reactions in porous media. Adv Water Resour 32:1664–1673
- Battiato I, Tartakovsky DM, Tartakovsky AM, Scheibe TD (2011) Hybrid models of reactive transport in porous and fractured media. Adv Water Resour 34:1140–1150
- Benson DA, Meerschaert MM (2008) Simulation of chemical reaction via particle tracking: Diffusion-limited versus thermodynamic rate-limited regimes. Water Resour Res 44, no. 12, doi: 10.1029/2008WR007111
- Berkowitz B, Ewing RP (1998) Percolation theory and network modeling applications in soil physics. Surv Geophys 19:23–72
- Berkowitz B, Naumann C, Smith L (1994) Mass transfer at fracture intersections: An evaluation of mixing models. Water Resour Res 30:1765–1773
- Bernardi C, Maday Y, Patera AT (1994) A new nonconforming approach to domain decomposition:the mortar element method. *In:*Brezis H, Lions JL (eds) Nonlinear Partial Differential Equations and their Applications. Collège de France Seminar. Vol. XI. Harlow, Longman Scientific & Technical. (Pitman Research Notes in Mathematics Series 299)
- Bijeljic B, Blunt MJ (2006) A Physically-Based Description of Dispersion in Porous Media. In:SPE Annual Tech Conf and Exhibition. Society of Petroleum Engineers, SPE-102869-MS
- Bijeljic B, Blunt MJ (2007) Pore-scale modeling of transverse dispersion in porous media. Water Resour Res 43:W12S11, doi: 10.1029/2006WR005700
- Bijeljic B, Muggeridge AH, Blunt MJ (2004) Pore-scale modeling of longitudinal dispersion. Water Resour Res 40:W11501, doi: 10.1029/2004WR003567
- Blunt MJ (2001) Flow in porous media—pore-network models and multiphase flow. Curr Opin Colloid In 6:197–207
- Boso F, Battiato I (2013) Homogenizability conditions for multicomponent reactive transport. Adv Water Resour 62:254–265
- Bruderer C, Bernabé Y (2001) Network modeling of dispersion: Transition from Taylor dispersion in homogeneous networks to mechanical dispersion in very heterogeneous ones. Water Resour Res 37:897–908
- Bryant SL, Blunt MJ (1992) Prediction of relative permeability in simple porous media. Phys Rev A 46:2004–2011, doi:http://dx.doi.org/10.1103/PhysRevA.46.2004
- Bryant SL, Mellor DW, Cade CA (1993a) Physically representative network models of transport in porous media. AIChE J 39:387–396
- Bryant SL, Cade CA, Mellor DW (1993b) Permeability prediction from geologic models. AAPG Bull 77:1338–1350
- Bryntesson LM (2002) Pore network modelling of the behaviour of a solute in chromatography media:transient and steady-state diffusion properties. J Chromatogr A 945:103–115
- Celia MA, Reeves PC, Ferrand LA (1995) Recent advances in pore scale models for multiphase flow in porous media. Rev Geophys 33, S2:1049–1057
- Chen S, Doolen GD (1998) Lattice Boltzmann method for fluid flows. Annu Rev Fluid Mech 30:329–364
- Chen L, Kang Q, Robinson BA, He YL, Tao WQ (2013) Pore-scale modeling of multiphase reactive transport with phase transitions and dissolution–precipitation processes in closed systems. Phys Rev E 87:043306

- Chen L, Kang Q, Tang Q, Robinson BA, He YL, Tao WQ (2014) Pore-scale simulation of multicomponent multiphase reactive transport with dissolution and precipitation. arXiv preprint arXiv:1410.3464
- Chu J, Engquist B, Prodanovič M, Tsai R (2012) A multiscale method coupling network and continuum models in porous media I:Steady-state single phase flow. Multiscale Model Simul 10:515–549
- Chu J, Engquist B, Prodanovič M, Tsai R (2013) A multiscale method coupling network and continuum models in porous media II—Single-and two-phase flows. *In:* Advances in Applied Mathematics, Modeling, and Computational Science, Springer, p 161–185
- Coelho D, Thovert JF, Adler PM (1997) Geometrical and transport properties of random packings of spheres and aspherical particles. Phys Rev E 55, doi:http://dx.doi.org/10.1103/PhysRevE.55.1959
- de Arcangelis L, Koplik J, Redner S, Wilkinson D (1986) Hydrodynamic dispersion in network models of porous media. Phys Rev Lett 57:996
- Dentz M, Le Borgne T, Englert A, Bijeljic B (2011) Mixing, spreading and reaction in heterogeneous media: A brief review. J Contam Hydrol 120:1–17
- Dong H, Blunt MJ (2009) Pore-network extraction from micro-computerized-tomography images. Phys Rev E 80:036307
- Edwards DA, Shapiro M, Brenner H, Shapira M (1991) Dispersion of inert solutes in spatially periodic, twodimensional model porous media. Transport Porous Med 6:337–358
- Ellis JS, Ebrahimi A, Bazylak A (2011) Characterization of a Two-Phase Pore-Scale Network Model of Supercritical Carbon Dioxide Transport Within Brine-Filled Porous Media. *In:* ASME 2011 5th International Conference on Energy Sustainability, p 1307–1314
- Fatt I (1956a) The network model of porous media I. Capillary pressure characteristics. Trans Am I Min Met Eng 207:144–159
- Fatt I (1956b) The network model of porous media. 2. Dynamic properties of a single size tube network. Trans Am I Min Met Eng 207:160–163
- Fatt I (1956c) The network model of porous media III. Dynamic properties of networks with tube radius distribution. Trans Am I Min Eng 207:164–181
- Finney JL (1968) Random packings and the structure of the liquid state. University of London (United Kingdom), PhD dissertation
- Fischer U, Celia MA (1999) Prediction of relative and absolute permeabilities for gas and water from soil water retention curves using a pore-scale network model. Water Resour Res 35:1089–1100
- Ganis B, Juntunen M, Pencheva G, Wheeler MF, Yotov I (2012) A Global Jacobian Method for Simultaneous Solution of Mortar and Subdomain Variables in Nonlinear Porous Media Flow. ICES REPORT 12–46, The Institute for Computational Engineering and Sciences, The University of Texas at Austin
- Gillespie DT (1977) Exact stochastic simulation of coupled chemical reactions. J Phys Chem 81:2340–2361
- Hoefner ML, Fogler HS (1988) Pore evolution and channel formation during flow and reaction in porous media. AIChE J 34:45–54
- Iltis GC, Armstrong RT, Jansik DP, Wood BD, Wildenschild D (2011) Imaging biofilm architecture within porous media using synchrotron-based X-ray computed microtomography. Water Resour Res 47:W02601, doi: 10.1029/2010WR009410
- Jha RK, Bryant SL, Lake LW (2011) Effect of diffusion on dispersion. SPE J 16:65-77
- Joekar-Niasar V (2010) The immischibles:Capillarity effects in porous media, pore-network modeling. Geologica Ultraiectina 318
- Joekar-Niasar V, Hassanizadeh SM, Dahle HK (2010) Non-equilibrium effects in capillarity and interfacial area in two-phase flow:dynamic pore-network modelling. J Fluid Mech 655:38-71
- Johnson J, Brown S, Stockman H (2006) Fluid flow and mixing in rough-walled fracture intersections. J Geophys Res:Solid Earth 111:B12206, doi: 10.1029/2005JB004087
- Kang Q, Lichtner PC, Viswanathan HS, Abdel-Fattah AI (2010) Pore scale modeling of reactive transport involved in geologic CO2 sequestration. Transport Porous Med 82:197–213
- Karadimitriou NK, Hassanizadeh SM (2012) A review of micromodels and their use in two-phase flow studies. Vadose Zone J 11, doi:10.2136/vzj2011.0072.
- Kechagia PE, Tsimpanogiannis IN, Yortsos YC, Lichtner PC (2002) On the upscaling of reaction-transport processes in porous media with fast or finite kinetics. Chem Eng Sci 57:2565–2577
- Kim D, Lindquist WB (2012) Dependence of pore-to-core up-scaled reaction rate on flow rate in porous media. Transport Porous Med 94:555–569
- Kim D, Peters CA, Lindquist WB (2011) Upscaling geochemical reaction rates accompanying acidic CO₂saturated brine flow in sandstone aquifers. Water Resour Res 47:W01505, doi: 10.1029/2010WR009472
- Koplik J, Lasseter TJ (1985) Two-phase flow in random network models of porous media. Soc Petrol Eng J 25:89–100
- Koplik, J, Redner S, Wilkinson D (1988) Transport and dispersion in random networks with percolation disorder. Phys Rev A 37:2619-2636

- Li L, Peters CA, Celia MA (2006) Upscaling geochemical reaction rates using pore-scale network modeling. Adv Water Resour 29:1351–1370
- Li L, Peters CA, Celia MA (2007a) Applicability of averaged concentrations in determining geochemical reaction rates in heterogeneous porous media. Am J Sci 307:1146–1166
- Li L, Peters CA, Celia MA (2007b) Effects of mineral spatial distribution on reaction rates in porous media. Water Resour Res 43:W01419, doi: 10.1029/2005WR004848
- Li L, Steefel CI, Yang L (2008) Scale dependence of mineral dissolution rates within single pores and fractures. Geochim Cosmochim Acta 72:360–377
- Lindquist WB, Lee SM, Coker DA, Jones KW, Spanne P (1996) Medial axis analysis of void structure in three-dimensional tomographic images of porous media. J Geophys Res:Solid Earth 101(B4):8297–8310
- Lindquist WB, Venkatarangan A (1999) Investigating 3D geometry of porous media from high resolution images. Phys Chem Earth Pt A 24:593–599
- Lopez X, Valvatne PH, Blunt MJ (2003) Predictive network modeling of single-phase non-Newtonian flow in porous media. J Colloid Interface Sci 264:256–265
- Lu C, Yortsos YC (2005) Dynamics of forward filtration combustion at the pore-network level. AIChE J 51:1279–1296
- Maier RS, Kroll DM, Bernard RS, Howington SE, Peters JF, Davis HT (2000) Pore-scale simulation of dispersion. Phys Fluids 12:2065–2079
- Maier RS, Kroll DM, Bernard RS, Howington SE, Peters JF, Davis HT (2003) Hydrodynamic dispersion in confined packed beds. Phys Fluids 15:3795–3815
- Martins AA, Laranjeira PE, Braga CH, Mata TM (2009) Modeling of transport phenomena in porous media using network models. *In:* Progress in Porous Media Research, ISBN:978-1-60692-435-8
- Meakin P, Tartakovsky AM (2009) Modeling and simulation of pore-scale multiphase fluid flow and reactive transport in fractured and porous media. Rev Geophys 47: RG3002, doi: 10.1029/2008RG000263
- Mehmani Y (2014) Modeling of single-phase flow and solute transport across scales. PhD dissertation, University of Texas at Austin
- Mehmani Y, Balhoff MT (2014) Bridging from pore to continuum: A hybrid mortar domain decomposition framework for subsurface flow and transport. Multiscale Model Simul 12:667–693
- Mehmani A, Prodanovič M (2014a) The effect of microporosity on transport properties in porous media. Adv Water Resour 63:104–119
- Mehmani A, Prodanovič M (2014b) The application of sorption hysteresis in nano-petrophysics using multiscale multiphysics network models. Int J Coal Geol 128:96–108
- Mehmani Y, Sun T, Balhoff MT, Eichhubl P, Bryant SL (2012) Multiblock pore-scale modeling and upscaling of reactive transport:application to carbon sequestration. Transport Porous Med 95:305–326
- Mehmani A, Prodanovič M, Javadpour F (2013) Multiscale, multiphysics network modeling of shale matrix gas flows. Transport Porous Med 99:377–390
- Mehmani Y, Oostrom M, Balhoff MT (2014) A streamline splitting pore-network approach for computationally inexpensive and accurate simulation of transport in porous media. Water Resour Res 50:2488–2517
- Moebius F, Or D (2014) Inertial forces affect fluid front displacement dynamics in a pore-throat network model. Phys Rev E 90:023019
- Mohanty KK, Davis HT, Scriven LE (1987) Physics of oil entrapment in water-wet rock. SPE Reservoir Eng 2:113–128
- Molins S (2015) Reactive interfaces in direct numerical simulation of pore scale processes. Rev Mineral Geochem 80:461–481
- Molins S, Trebotich D, Steefel CI, Shen C (2012) An investigation of the effect of pore scale flow on average geochemical reaction rates using direct numerical simulation. Water Resour Res 48:W03527, doi: 10.1029/2011WR011404
- Molins S, Trebotich D, Yang L, Ajo-Franklin JB, Ligocki TJ, Shen C, Steefel CI (2014) Pore-scale controls on calcite dissolution rates from flow-through laboratory and numerical experiments. Env Sci Tech 48:7453–7460, doi:10.1021/es5013438
- Monaghan JJ (1992) Smoothed particle hydrodynamics. Annu Rev Astron Astrophys 30:543-574
- Mostaghimi P, Bijeljic B, Blunt MJ (2012) Simulation of flow and dispersion on pore-space images. *In:* SPE Annual Technical Conference and Exhibition, Italy. SPE 135261
- Nogues JP, Fitts JP, Celia MA, Peters CA (2013) Permeability evolution due to dissolution and precipitation of carbonates using reactive transport modeling in pore networks. Water Resour Res 49:6006–6021
- Noiriel C, Bernard D, Gouze P, Thibault X (2005) Hydraulic properties and microgeometry evolution accompanying limestone dissolution by acidic water. Oil Gas Sci Tech 60:177–192
- Noiriel C (2015) Resolving time-dependent evolution of pore-scale structure, permeability and reactivity using X-ray microtomography. Rev Mineral Geochem 80:247–285

- Oostrom M, Mehmani Y, Romero-Gomez P, Tang Y, Liu H, Yoon H, Kang Q, Joekar-Niasar V, Balhoff MT, Dewers T, Tartakovsky GD, Leist EA, Hess NJ, Perkins WA, Rakowski CL, Richmond MC, Serkowski JA, Werth CJ, Valocchi AJ, Wietsma TW, Zhang C (2014) Pore-scale and continuum simulations of solute transport micromodel benchmark experiments. Comput Geosci 1–23, doi 10.1007/s10596-014-9424-0
- Øren PE, Bakke S, Arntzen OJ (1998) Extending predictive capabilities to network models." SPE J 3:324– 336. SPE 38880
- Øren PE, Bakke S (2002) Process based reconstruction of sandstones and prediction of transport properties. Transport Porous Med 46:311–343
- Ovaysi S, Piri M (2011) Pore-scale modeling of dispersion in disordered porous media. J Contam Hydrol 124:68–81
- Park YJ, Lee KK (1999) Analytical solutions for solute transfer characteristics at continuous fracture junctions. Water Resour Res 35:1531–1537
- Park YJ, Lee KK, Berkowitz B (2001a) Effects of junction transfer characteristics on transport in fracture networks. Water Resour Res 37:909–923
- Park YJ, Dreuzy JR, Lee KK, Berkowitz B (2001b) Transport and intersection mixing in random fracture networks with power law length distributions. Water Resour Res 37:2493–2501
- Parmigiani A, Huber C, Bachmann O, Chopard B (2011) Pore-scale mass and reactant transport in multiphase porous media flows. J Fluid Mech 686:40–76, doi:10.1017/jfm.2011.268
- Patzek TW (2001) Verification of a complete pore network simulator of drainage and imbibition. SPE J 6:144–156. SPE-71310-PA
- Payatakes AC, Ng KM, Flumerfelt RW (1980) Oil ganglion dynamics during immiscible displacement:Model formulation. AIChE J 26:430–443
- Peszyńska M, Wheeler MF, Yotov I (2002) Mortar upscaling for multiphase flow in porous media. Computat Geosci 6:73–100
- Picard G, Frey K (2007) Method for modeling transport of particles in realistic porous networks: Application to the computation of NMR flow propagators. Phys Rev E 75:066311
- Piri M, Blunt MJ (2005) Three-dimensional mixed-wet random pore-scale network modeling of two-and three-phase flow in porous media. I. Model description. Phys Rev E 71:026301
- Prodanovič M, Lindquist WB, Seright RS (2006) Porous structure and fluid partitioning in polyethylene cores from 3D X-ray microtomographic imaging. J Colloid Interface Sci 298:282–297
- Prodanovič M, Mehmani A, Sheppard AP (2014) Imaged-based multiscale network modelling of microporosity in carbonates. Geological Society, London, Special Publications 406:SP406-9
- Raoof A, Nick HM, Hassanizadeh SM, Spiers CJ (2013) PoreFlow: A complex pore-network model for simulation of reactive transport in variably saturated porous media. Comput Geosci 61:160–174
- Rhodes ME, Blunt MJ (2006) An exact particle tracking algorithm for advective-dispersive transport in networks with complete mixing at nodes. Water Resour Res 42: W04501, doi: 10.1029/2005WR004504
- Rhodes ME, Bijeljic B, Blunt MJ (2009) A rigorous pore-to-field-scale simulation method for single-phase flow based on continuous-time random walks. SPE J 14:88–94
- Roubinet D, Tartakovsky DM (2013) Hybrid modeling of heterogeneous geochemical reactions in fractured porous media. Water Resour Res 49:7945–7956
- Sahimi M, Hughes BD, Scriven LE, Davis HT (1986) Dispersion in flow through porous media—I. Onephase flow. Chem Eng Sci 41:2103–2122
- Scheibe TD, Tartakovsky AM, Tartakovsky DM, Redden GD, Meakin P (2007) Hybrid numerical methods for multiscale simulations of subsurface biogeochemical processes. J Phys:Conf Ser 78:012063. IOP Publishing. Doi:10.1088/1742-6596/78/1/012063
- Scheibe TD, Murphy EM, Chen X, Rice AK, Carroll KC, Palmer BJ, Tartakovsky AM, Battiato I, Wood BD (2015) An analysis platform for multiscale hydrogeologic modeling with emphasis on hybrid multiscale methods. Groundwater 53:38–56
- Sheng Q, Thompson KE (2013) Dynamic coupling of pore-scale and reservoir-scale models for multiphase flow. Water Resour Res 49:5973–5988
- Sheppard AP, Sok RM, Averdunk H, Robins VB, Ghous A (2006) Analysis of rock microstructure using high-resolution X-ray tomography. *In*:Proceedings of the International Symposium of the Society of Core Analysts, Trondheim, Norway. SCA2006-26
- Silin D, Patzek T (2006) Pore space morphology analysis using maximal inscribed spheres. Physica A 371:336–360
- Sorbie KS, Clifford PJ (1991) The inclusion of molecular diffusion effects in the network modelling of hydrodynamic dispersion in porous media. Chem Eng Sci 46:2525–2542
- Steefel CI, MacQuarrie KTB (1996) Approaches to modeling of reactive transport in porous media. Rev Mineral Geochem 34:85–129

- Steefel CI, DePaolo DJ, Lichtner PC (2005) Reactive transport modeling: An essential tool and a new research approach for the Earth sciences. Earth Planet Sci Lett 240:539–558
- Steefel CI, Molins S, Trebotich D (2013) Pore scale processes associated with subsurface CO₂ injection and sequestration. Rev Mineral Geochem 77:259–303
- Stehfest H (1970) Algorithm 368: Numerical inversion of Laplace transforms [D5]. Commun ACM 13:47–49 Suchomel BJ, Chen BM, Allen MB (1998a) Macroscale properties of porous media from a network model of biofilm processes. Transport Porous Med 31:39–66
- Suchomel BJ, Chen BM, Allen MB (1998b) Network model of flow, transport and biofilm effects in porous media. Transport Porous Med 30:1–23
- Sun T, Mehmani Y, Bhagmane J, Balhoff MT (2012a) Pore to continuum upscaling of permeability in heterogeneous porous media using mortars. Int J Oil Gas Coal Tech 5:249–266
- Sun T, Mehmani Y, Balhoff MT (2012b) Hybrid multiscale modeling through direct substitution of porescale models into near-well reservoir simulators. Energy & Fuels 26:5828–5836
- Tartakovsky AM, Tartakovsky DM, Scheibe TD, Meakin P (2008a) Hybrid simulations of reaction-diffusion systems in porous media. SIAM J Sci Comp 30:2799–2816
- Tartakovsky AM, Redden G, Lichtner PC, Scheibe TD, Meakin P (2008b) Mixing-induced precipitation:Experimental study and multiscale numerical analysis. Water Resour Res 44:W06S04 doi: 10.1029/2006WR005725
- Tartakovsky AM, Tartakovsky DM, Meakin P (2008c) Stochastic Langevin model for flow and transport in porous media. Phys Rev Lett 101:044502
- Thane CG (2006) Geometry and topology of model sediments and their influence on sediment properties, Thesis, University of Texas at Austin
- Thauvin F, Mohanty KK (1998) Network modeling of non-Darcy flow through porous media. Transp Porous Med 31:19–37
- Thompson KE, Willson CS, White CD, Nyman S, Bhattacharya JP, Reed AH (2008) Application of a new grain-based reconstruction algorithm to microtomography images for quantitative characterization and flow modeling. No. NRL/JA/7430-07-13. Naval Research Lab Stennis Space Center Ms Marine Geoacoustics Div
- Thovert JF, Salles J, Adler PM (1993) Computerized characterization of the geometry of real porous media:their discretization, analysis and interpretation. J Microsc 170:65–79
- Tomin P, Lunati I (2013) Hybrid Multiscale Finite Volume method for two-phase flow in porous media. J Comput Phys 250:293–307
- Trebotich D, Adams MF, Molins S, Steefel CI, Shen C (2014) High resolution simulation of pore scale reactive transport processes associated with carbon sequestration. Comput Sci Eng, 10.1109/MCSE.2014.77
- van Milligen BP, Bons PD (2014) Simplified numerical model for clarifying scaling behavior in the intermediate dispersion regime in homogeneous porous media. Comput Phys Comm 185:3291–3301. Doi:10.1016/j.cpc.2014.09.006
- Varloteaux C, Békri S, Adler PM (2013) Pore network modelling to determine the transport properties in presence of a reactive fluid: From pore to reservoir scale. Adv Water Resour 53:87–100
- Vogel HJ (2000) A numerical experiment on pore size, pore connectivity, water retention, permeability, and solute transport using network models. Eur J Soil Sci 51:99–105
- Wendt JF (2009) Computational Fluid Dynamics. Third Edition, Springer, Berlin
- Werth CJ, Zhang C, Brusseau ML, Oostrom M, Baumann T (2010) A review of non-invasive imaging methods and applications in contaminant hydrogeology research. J Contam Hydrol 113:1–24
- Wildenschild D, Sheppard AP (2013) X-ray imaging and analysis techniques for quantifying pore-scale structure and processes in subsurface porous medium systems. Adv Water Resour 51:217–246
- Yang X, Scheibe TD, Richmond MC, Perkins WA, Vogt SJ, Codd SL, Seymour JD, McKinley MI (2013) Direct numerical simulation of pore-scale flow in a bead pack:Comparison with magnetic resonance imaging observations. Adv Water Resour 54:228–241
- Yoon H, Valocchi AJ, Werth CJ, Dewers T (2012) Pore-scale simulation of mixing-induced calcium carbonate precipitation and dissolution in a microfluidic pore network. Water Resour Res 48: W02524 doi: 10.1029/2011WR011192
- Yoon H, Kang Q, Valocchi A (2015) Lattice Boltzmann-based approaches for pore-scale reactive transport. Rev Mineral Geochem 80:393–431
- Zaretskiy Y, Geiger S, Sorbie K, Förster M (2010) Efficient flow and transport simulations in reconstructed 3D pore geometries. Adv Water Resour 33:1508–1516