1 **Redox-Induced Nucleation and Growth of Goethite on Synthetic Hematite Nanoparticles**

2 Jeanette L. Voelz [a], William A. Arnold [b], R. Lee Penn [a,*]   **(Revision 1)**

3 [a]*Department of Chemistry, University of Minnesota, Minneapolis, Minnesota, 55455*

4 [b]*Department of Civil, Environmental, and Geo- Engineering, University of Minnesota, Minneapolis,*

5 *Minnesota, 55455*

6 *Corresponding author. Present address: 139 Smith Hall, 207 Pleasant Street SE, Minneapolis, MN

7 55455, USA. Fax: +1 (612) 626 7541. *E-mail address:* rleepenn@umn.edu (R. L. Penn).

8

9 **Abstract**

10

11 The iron (oxyhydr)oxides hematite (α-Fe$_2$O$_3$) and goethite (α-FeOOH) are natural and reactive

12 minerals common in soils and sediments, and their adsorption of Fe(II) produces reactive surface

13 sites that facilitate reduction of oxidized environmental pollutants. Single-exposure experiments

14 with 4-chloronitrobenzene showed that hematite is more reactive than goethite, when normalized

15 by surface area loading. Interestingly, the product of Fe(II) oxidation is a mixture of goethite

16 and hematite, and the goethite to hematite ratio depends on the distribution of Fe(II) activated

17 surface sites, which is a function of aqueous Fe(II) concentration, surface area loading, and pH.

18 More goethite is produced under conditions of higher Fe(II), lower surface area loading, and

19 higher pH. Recurrent-exposure experiments showed a substantial decrease in reaction rate after

20 one to three exposures, a trend suggestive of reaction contributions from the increasing goethite

21 surface area over time. Using known atomic surface geometry for goethite and hematite, the

22 hematite {012} facet is proposed as the site of primary mineral growth with goethite {021} at the

23 interface between the two minerals. These results have implications in contaminant fate

24   modeling, where the mineral phases present in the environment, the minerals likely to form, and

25   the surrounding aqueous conditions all have an impact on contaminant reaction rate.

26

27   **Introduction**

28

29   Iron is one of the most abundant elements on Earth, and its oxides, oxyhydroxides, and

30   hydroxides (collectively termed 'iron oxides') are currently tabulated at seventeen recognized

31   crystal structures (Faivre 2016). Iron oxide minerals form in a range of mild to extreme

32   environmental conditions, with phases produced dependent on temperature, pressure, pH,

33   aqueous ion concentrations, hydration state, and oxygen fugacity (Schwertmann and Murad

34   1983; Cornell and Schwertmann 1996; Lagroix et al. 2016).

35   Anthropogenic influences, however, rapidly alter minerals, and the geological impacts of

36   industrial activities like mining (Chopard et al. 2017; Valencia-Avellan et al. 2017) and

37   agriculture (Maqueda et al. 2017) are still under investigation. The release of highly oxidized

38   chemicals into reduced groundwaters facilitates redox reactions that not only transform the

39   contaminants (Elsner et al., 2004a; Howard and Muir, 2013) but alter mineral surfaces as well

40   (Chun et al. 2006; Strehlau et al. 2016). Minerals are recognized as important tools for

41   environmental remediation via natural attenuation (Pecher et al., 2002; Elsner, et al., 2004b),

42   permeable reactive barriers (Scherer et al. 2000), or *in-situ* redox manipulation (Boparai et al.

43   2008), but alterations from inadvertent anthropogenic influences must also be considered so that

44   they are recognized, rather than mistaken for natural processes.

45   The specific adsorption of Fe(II) to iron oxide surfaces is known to occur through bidentate

46   complexes with contiguous singly-coordinated hydroxyl group pairs (Wehrli et al. 1989; Barron

2

47 and Torrent 1996). The subsequent electron transfer to the bulk oxide results in the formation of

48 new Fe(III) surface structures (Tanwar et al. 2008; Catalano et al. 2010) and activates these

49 minerals for participation in redox reactions. This electron transfer process has been

50 demonstrated for goethite and hematite using Mössbauer spectroscopy (Williams and Scherer

51 2004; Larese-Casanova and Scherer 2007; Rosso et al. 2010) and calculated in detail for

52 hematite using molecular orbital calculations (Rosso et al. 2003). Many highly oxidized

53 environmental contaminants, such as nitroaromatic compounds, are readily reduced when they

54 encounter these Fe(II)-activated surfaces (Hofstetter et al. 1999; Dunnivant et al. 1992; Klausen

55 et al. 1995; Ju and Parales 2010). Measuring the concurrent oxidative mineral growth provides

56 important information about which surfaces were readily accessible to and most reactive with

57 aqueous Fe(II) (Chun et al. 2006; Vindedahl et al. 2015; Strehlau et al. 2016).

58   It has been proposed that Fe(II) adsorption on hematite leads to magnetite precipitation (Jeon

59 et al. 2001, 2003), but this proposal has not been supported in other works (Pedersen et al. 2005;

60 Tanwar et al. 2008). Rather, observations of Fe(II) adsorption on hematite {012} by crystal

61 truncation rod diffraction indicate that the oxidized surficial Fe(III) have Fe-O bond lengths of

62 1.93 – 2.18 Å, which is in agreement with Fe-O bond lengths in the bulk lattice of both goethite

63 and hematite but not magnetite (Yang et al. 2006; Tanwar et al. 2008). These reports of Fe(III)

64 surface structures resulting from the adsorption of Fe(II) on hematite are limited by the

65 adsorption capacity of the mineral surface (Tanwar et al. 2008).

66   To further investigate the evolving mineralogy requires the addition of another electron-

67 accepting species to oxidize Fe(II) beyond the adsorption capacity of the underlying hematite.

68 For example, the model contaminant 4-chloronitrobenzene (4-ClNB) efficiently participates in

69 redox reactions in systems containing iron oxides and Fe(II) (Elsner et al. 2004b; Vindedahl et

3

70   al. 2015; Strehlau et al. 2017). Oxidation of Fe(II) in suspensions of hematite may result in (1)

71   hematite growth on existing hematite faces, (2) homogeneous nucleation of new particles, or (3)

72   heterogeneous nucleation of new phases on existing hematite surfaces. Oxidative growth on

73   goethite and hematite surfaces after the addition of Fe(II) and reaction with nitrobenzenes has

74   been characterized using Mössbauer spectroscopy (Larese-Casanova et al. 2012). On hematite,

75   the deposition of 4-6 monolayers of hematite was observed, followed by the formation of a new

76   phase identifiable as goethite, which comprised 25 % and 95 % of the growth on hematite

77   needles and hexagonal platelets, respectively. Changes to solution conditions (pH, buffer

78   identity, ionic strength, and Fe(II) concentration) were not shown to have a profound effect on

79   the monolayer onset of goethite formation on hematite, suggesting that goethite formation is

80   instead a function of thermodynamic factors.

81   Evolving surface features resulting from oxidative growth (i.e. particle size, morphology,

82   surface roughness, or phase composition) affect the availability and accessibility of Fe(II)-

83   reactive sites. The comparison of particle dimensions via electron microscopy before and after

84   reaction reveal the locations of oxidative mineral growth, which may indicate the relative Fe(II)

85   reactivities of various facets (Chun et al. 2006; Vindedahl et al. 2015; Strehlau et al. 2016). For

86   example, where goethite particle tips became roughened in appearance after redox, the surfaces

87   could no longer be well-described as terminated by {021} facets that were prominent prior to

88   reaction, rather they compose a mixture of crystal surfaces, which changes the ratio of surface

89   sites and, therefore, affects the reaction rate (Chun et al. 2006).

90   Recurrent-exposure reactions are a valuable tool for evaluating the evolving reactivity of

91   mineral surfaces. As an example, two studies investigating the evolving reactivity of goethite in

92   1 mM Fe(II) after five exposures of 100 μM 4-ClNB report decreasing reaction rates with each

4

93   exposure (Chun et al. 2006; Vindedahl et al. 2015). Both these studies, however, did not include

94   pH readjustment after each sequential spike, which is expected to affect reaction kinetics as

95   surface hydroxyl groups are more densely protonated at lower pH, therefore inhibiting the

96   adsorption of Fe(II) and limiting the available reaction sites. A recent study of goethite reactivity

97   under related conditions showed no significant change in rate across five to ten additions of 4-

98   ClNB when the pH was readjusted after each sequential spike, suggesting the resulting oxidative

99   mineral growth formed a balance between inhibitory surface roughening and increasing reactive

100   surface area (Strehlau et al. 2016).

101   Results from the evolving reactivity of iron oxide surfaces enable a deeper understanding of

102   how redox reactions at the water-mineral interface may proceed in the environment, particularly

103   in cases where minerals have undergone substantial reactions with anthropogenic species. Here,

104   we investigate oxidative mineral growth as a function of the availability of Fe(II)-activated

105   surface sites at the hematite surface. The reaction variables tested include pH, Fe(II)

106   concentration, reactive surface area, and particle morphology. Both single-exposure and

107   recurrent-exposure reactions are performed to assess the effects of solution conditions on

108   reactivity and to evaluate structural changes due to oxidative growth. Finally, a detailed

109   description of the observed epitaxial growth is proposed based on lattice measurements and

110   known geometric atomic configurations of goethite and hematite facets.

111

112   **Experimental**

113

114   Ultrapure water from a MilliPore Milli-Q Advantage A10 system was used for all synthesis,

115   dialysis, and experimental processes. All solutions of HCl and NaOH for pH adjustments were

116 prepared from concentrated solutions of HCl (BDH Aristar, 12.1 M, 36.5 – 38.0 %) and NaOH

117 (Fisher Scientific, 19 M, 50 % w/w). All experiments were performed in a Coy Laboratory

118 Products vinyl anaerobic glovebag with 5 % H_2, 95 % N_2 (Matheson) atmosphere.

119

120 **Hematite syntheses**

121

122   Rhombohedral and equidimensional hematite particles were synthesized using procedures

123 modified from Cornell and Schwertmann (1996). Rhombohedral particles (hereafter denoted as

124 'R' in tables and figures) were synthesized in a 1 L Pyrex glass bottle, where 1 L of 0.002 M

125 HNO_3 (ACS Reagent Grade) was heated to 98 °C in an oven and 8.08 g (0.02 M Fe) of

126 $Fe(NO_3)_3 \cdot 9H_2O$ (Sigma-Aldrich, 98 %) was added with vigorous stirring until the salt dissolved.

127 The bottle was capped with aluminum foil (to minimize evaporation but also prevent

128 pressurizing) then promptly returned to the 98 °C oven and heated for seven days. The foil never

129 had contact with the solution. The suspension was cooled to room temperature, the supernatant

130 decanted, and the concentrated suspension was dialyzed using cellulose dialysis membranes

131 (Spectra/Por® 7, 2 kD, 38 mm × 24 mm). Dialysis water was changed nine times, with at least

132 four hours between water changes. The suspension was transferred to a Nalgene bottle, sparged

133 with nitrogen (Matheson 99.99 %) for two hours, and stored in the anaerobic glovebag.

134 Equidimensional particles (hereafter denoted as 'E' in tables and figures) were synthesized in a 2

135 L Nalgene bottle, where 2 L of 0.002 M HCl (BDH Aristar 36.5-38.0 %) were heated to 98 °C in

136 an oven and 10.81 g (0.02 M Fe) of $FeCl_3 \cdot 6H_2O$ (Fisher 97.3 %) was added with vigorous

137 stirring until dissolved. Paralleling the rhombohedral synthesis, the bottle was capped with

138 aluminum foil, promptly returned to the 98 °C oven, and heated for ten days. After cooling to

6

139   room temperature, the suspension was concentrated, dialyzed, sparged, and stored in the

140   anaerobic glovebag.

141

142   **Initial particle characterization**

143

144   Following dialysis, the mass loading of each hematite stock suspension was determined as

145   described in the Supplemental Information Section I (Fig. S1). The dried particles from the mass

146   loading analysis were gently ground into a powder for XRD analysis. The mineral content of the

147   powder was analyzed using a PANalytical X'Pert PRO X-ray diffractometer equipped with a

148   cobalt source (1.7909 Å) and an X'Celerator detector. Both synthetic materials were analyzed

149   from 10-90 ° 2θ with an effective dwell time of 100 s per step and a step size of 0.0167 ° 2θ.

150   Particle dimensions were determined using calibrated images collected with an FEI Tecnai

151   T12 transmission electron microscope operating at 120 kV. TEM samples were prepared by

152   diluting one drop of stock suspension with methanol until the suspension appeared lightly red

153   (ca. 1000x), one drop of which was placed onto a Cu 200 mesh holey carbon grid and air-dried.

154   Particle dimensions are reported as the average ± standard deviation of 500 measurements with

155   approximately ten particles measured per image.

156   Surface area measurements were performed using N_2 adsorption by Brunauer-Emmett-Teller

157   (BET) analysis on a Quantachrome Autosorb iQ2-MP at 77 K with outgassing at 90 °C for 12 h.

158   Specific surface areas were determined using the adsorption branch from $P/P_o = 0.1$ to 0.28.

159

160   **Batch reactor preparation**

161

7

162   Batch reactors were prepared in 120 mL serum bottles capped with PTFE-lined rubber septa

163   and magnetically stirred with PTFE-coated stirbars at a rate of 500 rpm. All reactors contained a

164   buffer of 10 mM $NaHCO_3$ (Sigma-Aldrich 99.7 %) adjusted to pH 7.0 ± 0.1 with 1 M HCl.

165   Particle surface area loading varied from 0.007 m^2/mL – 0.043 m^2/mL, which corresponds to

166   mass loading of 0.100 to 0.600 mg/mL for rhombohedral particles and 0.383 to 1.149 mg/mL for

167   equidimensional particles. The 175 mM Fe(II) stock solution was prepared from oxygen-free

168   solid $FeCl_2 \cdot 9H_2O$ (Fisher 101.5 %) and 0.3 M HCl. Reactors contained Fe(II) concentrations of

169   0.25, 0.50, or 1.00 mM. Reactors were loaded with hematite, pH buffer, and Fe(II), and then pH

170   adjusted, as needed, to 7.0 or 6.5 ± 0.1 using 1 M HCl or 0.5 M NaOH.

171

172   **Single-exposure and recurrent-exposure reaction sets**

173

174   Reactors receiving one spike of 4-ClNB (Acros Organics 99 %) are termed single-exposure

175   reactions and were prepared in duplicate along with an additional control reactor to which no 4-

176   ClNB was added. All prepared reactors equilibrated for 21 hours prior to 4-ClNB addition. A

177   10 mM 4-ClNB stock solution prepared in methanol (Fisher Chemical 99.9 %), was added to the

178   reactors to achieve an initial concentration of one-tenth the reactor Fe(II) concentration.

179   Reactors receiving multiple additions of 4-ClNB were labeled recurrent-exposure reactions. For

180   these, three to five serum bottles were identically prepared under the desired reaction conditions.

181   One reactor functioned as a control, with no injection of 4-ClNB nor additional Fe(II). The

182   remaining reactors received additions of 4-ClNB and sacrificed for post-reaction characterization

183   after one, two, three, or six exposures. To perform the reaction sequence, the prepared reactors

184   equilibrated for 21 hours, and then the appropriate amount of 4-ClNB stock solution was added

8

185   (10 % of the initial Fe(II) concentration). The first reactor in each sequence was monitored by

186   HPLC to determine the 4-ClNB transformation rate. After 24 hours, the first reactor was tested

187   for the presence of 4-ClNB again by HPLC. If 4-ClNB was no longer detected, the Fe(II)

188   concentration was measured by UV-visible spectroscopy and the amount of Fe(II) needed to

189   return the reactors to the initial Fe(II) concentration was added to all remaining reactors. Finally,

190   the pH was adjusted to 7.0 ± 0.1 with 0.5 M NaOH. The remaining reactors equilibrated for 21

191   hours before the next addition of 4-ClNB, and the process was repeated. After determining the

192   4-ClNB reaction rate in a reactor, it was sacrificed for post-reaction characterization. The

193   control was sacrificed along with the last reactor in the set. For a schematic describing the

194   recurrent-exposure process, see Fig. S2.

195

196   **Reaction kinetics**

197

198   Reaction kinetics were monitored by quantifying 4-ClNB concentration using high-

199   performance liquid chromatography on an Agilent 1220 Infinity equipped with Zorbax SB-C18

200   4.6×150 mm, 5-µm column. The eluent was composed of a 30 : 70 mixture of 12 mM

201   ammonium acetate (Sigma-Aldrich 97 %) in 90 vol% water and 10 vol% acetonitrile adjusted to

202   pH 7.0 with ammonium hydroxide (Sigma-Aldrich 28 %) : pure acetonitrile. The flow rate was

203   0.7 mL/min, and the detection wavelength was set to 254 nm. Standards with a concentration

204   range of 5 – 100 µM 4-ClNB were made in methanol. For sampling, 0.5 mL of the reaction

205   mixture was filtered into autosampler vials using PALL Life Sciences Acrodisc® 13 mm syringe

206   filters with 0.2 µm nylon membrane starting 1 min after the 4-ClNB injection and continued until

207   approximately 80 % of the 4-ClNB had been reduced (1 – 3 hours).

9

208

Fe(II) adsorption

210

211   The concentration of Fe(II) in the reactors was determined via the ferrozine assay and UV-

212   visible spectroscopy on an Agilent 8453 equipped with tungsten and deuterium lamps. A 10 mM

213   ferrozine (3-(2-pyridyl)-5,6-diphenyl-1,2,4-triazine-p,p'-disulfonic acid monosodium salt

214   hydrate, Sigma-Aldrich 97 %) was prepared in water. To remove suspended particles, 0.8 mL of

215   the reaction mixture was filtered through PALL Life Sciences Acrodisc® 13 mm syringe filters

216   with 0.2 μm nylon membrane. Polypropylene cuvettes measuring 1 cm wide were prepared in

217   triplicate with 0.4 mL of 10 mM ferrozine, 3.4 mL water, and 0.2 mL of filtered reaction

218   mixture. Standards prepared with ferrozine, water, and 0.15 mM $FeCl_2$ (diluted from 175 mM

219   stock $FeCl_2$ solution) ranged 0.005-0.050 mM Fe(II). Absorption at 562 nm was recorded in

220   triplicate for each cuvette. The Fe(II) concentration in the reaction mixture was analyzed prior to

221   4-ClNB addition and after the reaction was complete.

222

223   **Post-reaction characterization**

224

225   Particles were collected from sacrificed reactors by centrifugation using an Eppendorf 5804

226   centrifuge programmed to run at 4427 rcf for four minutes. The reactor contents were first

227   concentrated in a 50 mL centrifuge tube using up to three rounds of centrifugation, where the

228   tube was only opened under anaerobic atmosphere to decant the supernatant. Once concentrated,

229   the particles were washed four times with 35 mL aliquots of deoxygenated water, again only

230   decanted under anaerobic atmosphere. After the final wash, the tubes were opened to ambient

231   atmosphere, the supernatant decanted, and the concentrated slurry transferred to weigh boats to

232   dry at room temperature in a fume hood. A portion of the dried particles was gently ground to

233   produce a uniform powder for XRD analysis. For TEM analysis, a minute quantity of unabraded

234   powder was added to 0.5 mL of methanol and sonicated for 30 seconds or until the suspension

235   appeared light red. For electron diffraction, the same suspension was sonicated for 15 minutes to

236   assist in separating the rods from the larger particles. To prepare the samples for TEM analysis,

237   one drop of the suspension was placed onto a Cu 200 mesh holey carbon grid, allowed to dry

238   overnight, and analyzed on an FEI Technai T12 transmission electron microscope operated at

239   120 kV.

240

241   **XRD calibration curve**

242

243   An X-ray diffraction calibration curve was used to quantify hematite and goethite present

244   after reaction. Calibration standards containing $1 - 50$ mass% goethite were prepared from

245   aqueous suspensions using mixtures of goethite (10.00 mg/mL) and rhombohedral hematite (3.93

246   mg/mL). Goethite (space group Pnma) was obtained from the starting material used for

247   experiments described in Penn et al. (2006), and the particles had average dimensions of 90 nm

248   by 10 nm. Two standards were prepared with equidimensional hematite (8.29 mg/mL) to test

249   comparability. All standards and reaction samples were analyzed from 18-32° 2θ with an

250   effective dwell time of 950 s per step and a step size of 0.0167° 2θ. The (110) reflection of

251   goethite (at 24° 2θ with the Co source) and (012) reflection of hematite (at 28° 2θ with the Co

252   source) were chosen because of the strong scattering intensity and baseline separation from other

11

253 goethite and hematite reflections. Origin (OriginLab Corporation 2016, version 93E) was used

254 to zero the spectral background, and the peaks were individually fitted using the PsdVoigt1

255 function to obtain integrated peak areas. Peak areas were multiplied by the goethite (110) and

256 hematite (012) structure factors squared to correct for relative peak intensities, which were

257 calculated to be 31.9 and 102.9, respectively. See Supplemental Section II for a detailed

258 description of these calculations.

259

260 **Results**

261

262 **Initial hematite particle characterization**

263

264 Hematite particles with two different morphologies were successfully synthesized. Collected

265 XRD patterns for the as-synthesized materials are both consistent with the hematite PDF 33-

266 0664 (Fig. S3). No additional phases were detected. The rhombohedral hematite particle

267 dimensions were 43 ± 8 nm by 29 ± 6 nm, and the nominally equidimensional hematite particles

268 had a diameter measuring 88 ± 13 nm. See Fig. S4 for graphics describing measurement

269 geometry and particle size distribution. Specific surface areas were determined using N_2

270 adsorption and were 72.3 m^2/g and 18.9 m^2/g for rhombohedral and equidimensional hematite,

271 respectively.

272 Based on TEM images and electron diffraction patterns, the rhombohedral hematite particles

273 are well described as rhombohedra with a slightly truncated pseudocubic morphology, which is a

274 hybrid of two morphologies described by Guo and Barnard (2011). Purely rhombohedral

275 hematite is bound only by {101} facets, while fully truncated pseudocubic morphologies have

12

276 sizeable {012} facets at the far tips of the rhombohedra. TEM images of the particles used here

277 (Fig. S5) show only minor {012} facets, giving the particles a mostly rhombohedral morphology.

278 The crystallographic structure of hematite dictates that the {101} and {012} planes are

279 orthogonal, which is consistent with the lattice fringe spacings observed in TEM images (Fig.

280 S5) as well as the observation that the {012} lattice fringes are perpendicular to the adjacent

281 {101} facets of the rhombohedron. For the equidimensional particles, tilt series confirm the

282 equidimensional morphology, rather than a plate-like structure (not shown).

283

284 **Oxidative mineral growth on hematite nanoparticles**

285 The products of the oxidative mineral growth are a mixture of goethite and hematite, and the

286 relative amount of each mineral was tracked as a function of Fe(II) concentration, surface area

287 loading, pH, and the morphology (rhombohedral vs equidimensional) of the hematite particle.

288 The reference condition, which was selected because it is consistent with the aqueous conditions

289 used in previous work with goethite (Vindedahl et al. 2015; Strehlau et al. 2016), used the

290 rhombohedral particles with surface area loading of 0.0072 m^2/mL (0.100 mg/mL mass loading)

291 and 1 mM Fe(II) at pH 7.0 with one exposure of 100 μM 4-ClNB, and is hereafter denoted R-r.

292 The post-reaction solid contained 15(2) % goethite by mass as determined by XRD.

293

294   **Growth as a function of reaction conditions.**

295 To quantitively compare the products of oxidative growth as a function of reaction

296 conditions, the stoichiometric maximum of goethite that could be produced by oxidative mineral

297 growth, expressed as a mass percent in the final solid after the reaction has gone to completion,

298 was calculated. For each reaction condition, the stoichiometric mass percent of goethite was

13

299   calculated using the following restrictions: (1) all six electrons needed for the reduction of 4-

300   ClNB are provided through Fe(II) oxidizing to Fe(III) and (2) all Fe(III) is incorporated into the

301   crystal structure as α-FeOOH. For example, the stoichiometric maximum for R-r is 35 mass%

302   goethite after the reaction has gone to completion. The stoichiometric maximum of goethite for

303   each reaction condition is reported in Table 1 as goethite max (mass%). The difference in

304   aqueous Fe(II) measured before and after reaction is stoichiometrically consistent with the Fe(II)

305   mass balance (Fig. S6).

306   Set a (Table 1) tested the effect of increased surface area loading for a single-exposure

307   reaction of 100 μM 4-ClNB with 1 mM Fe(II) at pH 7.0. When the surface area loading was

308   doubled (R-a1) as compared to the reference condition, the formation of goethite decreased

309   substantially, with only trace goethite (< 1 mass%) detected by XRD. In reactors prepared with

310   even higher surface area loading (i.e. R-a2, R-a3, and R-a4), no goethite was detected. TEM

311   micrographs of the post-reaction material support this result, with few rods observed in R-a1 and

312   no rods observed in post-reaction material from R-a2, R-a3, and R-a4 (Fig. S7). With high

313   surface area loading, it is possible that Fe(III) is incorporated as goethite layers too thin to be

314   detected either in TEM images or through XRD.   Indeed, a quick calculation assuming 37 nm

315   spherical particles yields approximately a 1 nm shell distributed over the entire particle surface.

316   As such, recurrent-exposure reactions were used to further probe high surface area loading

317   conditions, where the ratio of total oxidized Fe(II) to initial hematite mass is held constant.

318   Set b (Table 1) tested the effect of increased surface area loading in recurrent-exposure

319   reactions, with the cumulative stoichiometric maximum of goethite equal to R-r, at 35 mass%.

320   Sequence R-b1 contained three times the surface area loading as R-r and received three additions

321   of 100 μM 4-ClNB, and sequence R-b2 contained six times the surface area loading as R-r and

14

322 received six additions of 4-ClNB. In comparison to R-r, there was a decrease in the observed

323 goethite content: from 15(2) mass% for one exposure in R-r, to 9.3(8) mass% for three exposures

324 in R-b1, and 1.9(4) mass% for six exposures in R-b2 (Fig. 1).

325 Set c (Table 1) tested the effect of decreased pH by using an initial pH of 6.5 while matching

326 all other variables of the reference reaction. After reaction, only trace goethite was detected (< 1

327 mass%) by XRD. TEM micrographs of the R-c1 post-reaction material supports this result, with

328 few rods observed (Fig. S7).

329 Set d (Table 1) tested the effect of decreased aqueous Fe(II) concentration. Two recurrent-

330 exposure experiments were prepared; as compared to the reference, the Fe(II) and 4-ClNB

331 concentrations were halved and two exposures were performed in R-d1, and the concentrations

332 were quartered with four exposures performed in R-d2. For both sequences, the stoichiometric

333 maximum of goethite was 35 mass%. All reactors in R-d1 and R-d2 were analyzed by XRD for

334 mineral content, and the trends are presented in Fig. 2. At halved concentrations, only trace

335 goethite (<1 mass%) was observed after the first exposure while the vast majority of the

336 observed goethite formed during the second exposure, giving a total observed goethite in R-d1 of

337 9(1) mass%. At quartered concentrations, there was no detectable goethite formed after either

338 the first or second exposure, but 2.3(5) mass% was observed after the third exposure, and the

339 remaining goethite formed in the fourth exposure to total 6.2(8) mass% in R-d2. The data

340 presented in Fig. 2 show two main points: (1) as Fe(II) concentration decreases, less goethite is

341 formed, and (2) under recurrent-exposure conditions, goethite forms later in the exposure

342 sequence.

343 Set e (Table 1) tested how the aforementioned surface area loading results compare with

344 hematite of a different morphology, given that the exposed surfaces are quite different. The

15

345 single-exposure condition (E-e1) and recurrent-exposure condition (E-e2) were matched in

346 surface area loading to the rhombohedral reference R-r (0.0072 m^2/mL) and recurrent-exposure

347 R-b1 (0.0217 m^2/mL) as calculated from the BET measurements. Due to the larger particle size

348 of equidimensional hematite, and thus higher mass loading, the stoichiometric maximum of

349 goethite for each condition was 12 mass%. Condition E-e1 produced 5.2(7) mass% goethite and

350 condition E-e2 produced 1.9(4) mass% goethite.

351 Comparing reactors that have equivalent ratio of oxidized Fe(II) to surface area loading

352 demonstrates that the goethite mass percent formed on rhombohedral and equidimensional

353 hematite are both negatively correlated with an increase in surface area loading (Fig. 3). The

354 stoichiometric maxima of goethite for all reactors in Set b and Set e are 35 mass% and 12

355 mass%, respectively. The observed mass percent of goethite in each sample was lower than the

356 stoichiometric maximum, and some samples showed a trend: in the single-exposure experiments

357 with rhombohedral and equidimensional hematite (0.0072 m^2/mL surface area loading),

358 approximately 45 % of the stoichiometric maximum goethite formation was observed. In

359 recurrent-exposure experiments (0.0217 m^2/mL surface area loading), approximately 27 % of the

360 stoichiometric maximum goethite formed on the rhombohedral particles and 16 % formed on the

361 equidimensional particles.

362 Supporting the trends in oxidative mineral growth observed by XRD, TEM micrographs

363 show that the length and number of rod-like particles in the post-reaction material for both Set b

364 and Set e decreases as the surface area loading increases (Fig. 3). Images of samples containing

365 pure hematite (R, E of Fig. 3) exhibited no particles with the rod morphology characteristic of

366 goethite. There was no detectable goethite by XRD or TEM in any of the control reactors.

367

16

368   **Reaction kinetics**

369

370   The effects of solution conditions on hematite reactivity were studied in both single-exposure

371   and recurrent-exposure experiments. For comparison, literature values for goethite reactivity

372   (Table S1) were selected with the restrictions that mineral size and solution conditions were

373   similar to the conditions used in this study (Chun et al. 2006; Stemig et al. 2014; Vindedahl et al.

374   2015). All rates are normalized to reactor surface area loading (m^2/mL).

375

376   **Single-exposure reaction kinetics.**

377   The single-exposure reduction rate of 4-ClNB by Fe(II) in suspensions containing

378   rhombohedral and equidimensional hematite increases as the surface area loading increases, as

379   seen in Fig. 4a, which correlates with an increase in the number of available reactive sites. When

380   compared to reactions using goethite and 4-ClNB, rhombohedral and equidimensional hematite

381   react faster than goethite (Fig. 4a).

382

383   **Recurrent-exposure reaction kinetics.**

384   In the recurrent-exposure experiments of 4-ClNB on both rhombohedral and equidimensional

385   hematite (Fig. 4b), reaction rates drop substantially over the course of the exposure sequence.

386   For the first exposure, Fig. 4b shows again that as surface area loading increases, reaction rate

387   increases. For further 4-ClNB exposures, reaction rates progressively decrease with each

388   exposure. In the recurrent-exposure experiment containing R-b2, reactors were sacrificed for

389   determining the 4-ClNB reduction rate after one, three, and six exposures. The rate dropped

390   from 2.7(4) h^{-1} for the first exposure to 1.6(2) h^{-1} and 1.5(2) h^{-1} in the third and sixth exposure,

17

391   respectively. Further, in the recurrent-exposure experiment containing R-b1, the reaction rate

392   decreased from 2.0(3) h^{-1} in the first exposure to 1.2(2) h^{-1} after the third exposure. In both R-b1

393   and R-b2, the observed rates decrease to approximately 60 % of the initial rate. In the recurrent-

394   exposure experiment with equidimensional hematite, the rate decreased from 2.0(3) h^{-1} in the

395   first exposure to 0.7(1) h^{-1} then 0.9(1) h^{-1} after the second and third exposures, respectively.

396   These rates were both approximately 40-50 % of the initial rate.

397

398   **Post-reaction mineral identification by TEM**

399

400   TEM imaging shows that many of the rhombohedral particles have a 'kite-and-tail'

401   morphology after reaction with 4-ClNB, as seen in Fig. 5a. The tails, composed of newly-

402   formed material on hematite {012}, are referred to as rods within this work. Selected area

403   diffraction of extensively sonicated samples enabled facile collection of diffraction patterns of

404   the rods (Fig. 5b, left inset), which were consistent with the goethite crystal structure in all tested

405   samples (more examples given in Fig. S8). Lattice fringe measurements indicated the presence

406   of goethite in both rhombohedral and equidimensional post-reaction specimens, with the (110)

407   (Fig. 5b, right inset) and the (040) lattice planes (Fig. S8) aligned with the growth direction,

408   which is consistent with the known elongation direction of goethite.

409

410   **Discussion**

411

412   Contrary to what has been previously reported (Larese-Casanova et al. 2012), the mass

413   fraction of goethite in the post-reaction solids depended strongly on the reaction conditions: the

18

414   observed goethite mass percent in the post-reaction material decreased when surface area loading

415   increased, when initial pH decreased, and when the aqueous Fe(II) concentration decreased. The

416   difference between these results may be attributed to the conditions employed, which differed in

417   buffer identity (3-(*N*-morpholino)propanesulfonic acid), mass loading (>3×), Fe(II)

418   concentration (3×), and pH (>7.0). These results indicate that the heterogeneous nucleation of

419   new phases on the hematite surface is dependent on the distribution of Fe(II)-activated surface

420   sites in aqueous suspensions containing carbonate.

421

422   **Fe(II)-activated surface site geometry**

423   Structural considerations support the notion that the atomic surface geometry of hematite

424   {012} serves as an epitaxial template for goethite formation. The specific adsorption of Fe(II) is

425   known to occur at the contiguous singly-coordinated hydroxyl group pair sites (Wehrli et al.

426   1989; Barron and Torrent 1996); both goethite {021} and hematite {012} have approximately

427   four contiguous singly-coordinated hydroxyl group pairs per nm^2 (Table 2). The Fe atoms

428   directly below these contiguous singly-coordinated hydroxyl groups are positioned at equivalent

429   distances below the outermost oxygen atoms, at 0.178 and 0.177 nm in hematite and goethite,

430   respectively (Fig. 6). Additionally, these contiguous singly-coordinated hydroxyl group pairs

431   have the same spatial positioning above their coordinated Fe(III). Hematite {012} has

432   equivalent surface densities of triply- and singly-coordinated hydroxyl groups, at 7.3 hydroxyl

433   groups per nm^2, and no doubly-coordinated groups (Barron and Torrent 1996). The addition of

434   new Fe(III) at only one-half the hydroxyl groups on hematite {012} produces a layer of material

435   that has structural similarity to goethite, which has approximately one-half the Fe density of

436   hematite. Indeed, equilibrium experiments have shown a plateau in the maximum adsorption of

19

437   Fe(II) on hematite {012} where roughly half of the available binding sites were occupied

438   (Tanwar et al. 2008). These structural similarities are crystallographically consistent with the

439   observed transition in phase composition from hematite to goethite.

440   Rhombohedral hematite is one of the least thermodynamically favored hematite

441   morphologies (Guo and Barnard 2011). Because of the high free energy of formation and high

442   surface energies relative to goethite (Navrotsky et al. 2008; Guo and Barnard 2011), substantial

443   growth of the existing hematite crystal is unlikely in aqueous conditions at room temperature.

444   Additionally, goethite formation is favorable in aqueous systems containing carbonate (Cornell

445   and Schwertmann 1996). There are, however, substantial gaps in understanding this phase

446   transformation. Detailed thermodynamic calculations and modeling are needed, because reactive

447   surface structures are a function of Fe(II) concentration, equilibration time, peripheral ion

448   concentrations, pH, and extent of reaction with oxidized molecules, and the nature of evolving

449   reactivity calls for an iterative approach to such calculations.

450

451   **Reaction kinetics**

452   Faster reaction rates observed with higher surface area loading is expected because a greater

453   number of reaction sites allows more Fe(II) adsorption to occur and, thus, more electron transfer

454   to the bulk and more rapid 4-ClNB reduction. Differences in 4-ClNB reduction rates between

455   these results with hematite and literature rates for goethite (normalized for surface area loading)

456   arises from the differences in reactivity across accessible facets. For goethite, the majority of

457   oxidative mineral growth occurs on {021} rather than {110} in circumneutral pH conditions in

458   the absence of species that inhibit Fe(II) adsorption and/or the approach of the contaminant (e.g.

459   dissolved organic matter; Strehlau et al. 2016). The {021} of the goethite crystals comprise a

20

460   minor fraction of the accessible surface area. The faster rate observed for the reaction on

461   rhombohedral hematite could mean that either a greater fraction of the total accessible surface

462   area adsorbs Fe(II) or the secondary facets adsorb more Fe(II) than those on goethite, or both.

463   In recurrent-exposure experiments with hematite, the dramatic decrease in reaction rate

464   between the first and third exposures of 4-ClNB is a consequence of the changing distribution

465   and identity of accessible reactive sites. Initially, the surface is solely comprised of hematite.

466   Once goethite sites form, the reaction rate slows as hematite surface sites are blocked by the

467   goethite, and the more slowly reacting goethite surface sites comprise an increasing fraction of

468   the total accessible surface sites. This is shown in Fig. 4, where recurrent-exposure reaction rates

469   (Fig. 4b) decrease to a level fitting the trend of goethite rates (Fig. 4a). Growth formed through

470   further reactions would not be expected to have a substantial impact on rate, even as the goethite

471   surface area increases slightly. Indeed, previous studies have shown that goethite rate remains

472   relatively constant over pH-controlled recurrent-exposure reactions (Strehlau et al. 2016).

473

474   **Implications**

475

476   The evolving reactivity of hematite and goethite nanoparticles in anoxic aqueous

477   environments is important for applications in pollutant fate modeling. Projecting the fate and

478   transport of oxidized contaminants and their products in groundwater systems containing iron

479   oxides requires a detailed understanding not only of the chemical transformation of the

480   contaminant but also the characteristics of reactive sites on mineral surfaces and how those

481   characteristics change as reactions progress.

482   Specifically addressing the water-mineral interface, the reactivity of iron oxide surfaces is

483   sensitive to aqueous conditions such as Fe(II) concentration, pH, and availability of accessible

484   surface sites. The reactivity of the minerals, however, is determined by what surfaces are present

485   at a particular time. As the exposed surfaces grow or dissolve, the reactivity of the system

486   evolves and disordered or roughened surfaces lead to substantial changes in the kinetics of

487   contaminant reduction and concurrent oxidative mineral growth. Under conditions that favor

488   hematite production on hematite nanoparticles, reactivity is expected to be relatively unchanging.

489   Conditions that favor goethite production on hematite surfaces, however, could lead to vastly

490   different reaction rates; a hematite particle coated in a goethite shell would behave as goethite,

491   regardless of which mineral is present in larger mass.

492   With two mineral phases present, the most important parameter is accessible surface area of

493   the respective minerals. As reactions progress, however, this parameter is a moving target.

494   Ultimately, the aqueous conditions determine which mineral phase forms on the hematite

495   particles, and those conditions can be used to predict reactivity based on evolving mineralogy.

496   Illustrated by the results presented here for hematite particles, fate models for contaminated

497   ground water systems should consider not only initial mineral composition and surface area at

498   the time of contaminant exposure but also the minerals that may form as a result of surface

499   reactions to predict changes in pollutant transformation rates.

500

501   **Acknowledgements**

502

505 partial support from NSF through the MRSEC program. Thanks to Thomas E. Webber, who

506 performed the BET measurements.

507

508 **References**

509

510 Barron, V., and Torrent, J. (1996) Surface Hydroxyl Configuration of Various Crystal Faces of

511   Hematite and Goethite. Journal of Colloid and Interface Science, 177, 407–410.

512 Boparai, H.K., Comfort, S.D., Shea, P.J., and Szecsody, J.E. (2008) Remediating explosive-

513   contaminated groundwater by in situ redox manipulation (ISRM) of aquifer sediments.

514   Chemosphere, 71, 933–941.

515 Catalano, J.G., Fenter, P., Park, C., Zhang, Z., and Rosso, K.M. (2010) Structure and oxidation

516   state of hematite surfaces reacted with aqueous Fe(II) at acidic and neutral pH. Geochimica

517   et Cosmochimica Acta, 74, 1498–1512.

518 Chopard, A., Benzaazoua, M., Bouzahzah, H., Plante, B., and Marion, P. (2017) A contribution

519   to improve the calculation of the acid generating potential of mining wastes. Chemosphere,

520   175, 97–107.

521 Chun, C.L., Penn, R.L., and Arnold, W.A. (2006) Kinetic and Microscopic Studies of Reductive

522   Transformations of Organic Contaminants on Goethite. Environmental Science and

523   Technology, 40, 3299–3304.

524 Cornell, R.M., and Schwertmann, U. (1996) The Iron Oxides, 360 p. Wiley-VCH, Weinheim,

525   Germany.

526 Dunnivant, F.M., Schwarzenbach, R.P., and Macalady, D.L. (1992) Reduction of Substituted

527   Nitrobenzenes in Aqueous Solutions Containing Natural Organic Matter. Environmental

23

528   Science and Technology, 26, 2133–2141.

529   Elsner, M., Haderlein, S.B., Kellerhals, T., Luzi, S., Zwank, L., Angst, W., and Schwarzenbach,

530   R.P. (2004a) Mechanisms and Products of Surface-Mediated Reductive Dehalogenation of

531   Carbon Tetrachloride by Fe(II) on Goethite. Environmental Science and Technology, 38,

532   2058–2066.

533   Elsner, M., Schwarzenbach, R.P., and Haderlein, S.B. (2004b) Reactivity of Fe(II)-bearing

534   minerals toward reductive transformation of organic contaminants. Environmental Science

535   and Technology, 38, 799–807.

536   Faivre, D. (2016) Introduction. In D. Faivre, Ed., Iron Oxides: From Nature to Applications, p. 1-

537   5. Wiley - VCH, Weinheim, Germany.

538   Guo, H., and Barnard, A.S. (2011) Thermodynamic modelling of nanomorphologies of hematite

539   and goethite. Journal of Materials Chemistry, 21, 11566.

540   Hofstetter, T.B., Heijman, C.G., Haderlein, S.B., Holliger, C., and Schwarzenbach, R.P. (1999)

541   Complete reduction of TNT and other (poly)nitroaromatic compounds under iron-reducing

542   subsurface conditions. Environmental Science and Technology, 33, 1479–1487.

543   Howard, P.H., and Muir, D.C.G. (2013) Identifying New Persistent and Bioaccumulative

544   Organics Among Chemicals in Commerce. III: Byproducts, Impurities, and Transformation

545   Products. Environmental Science and Technology, 47, 5259–5266.

546   Jeon, B.H., Dempsey, B.A., Burgos, W.D., and Royer, R.A. (2001) Reactions of Ferrous Iron

547   with Hematite. Colloids and Surfaces A: Physicochemical and Engineering Aspects, 191,

548   41–55.

549   Jeon, B.H., Dempsey, B.A., and Burgos, W.D. (2003) Kinetics and mechanisms for reactions of

550   Fe(II) with iron(III) oxides. Environmental Science and Technology, 37, 3309–3315.

551   Ju, K.-S., and Parales, R.E. (2010) Nitroaromatic compounds, from synthesis to biodegradation.

552   Microbiology and Molecular Biology Reviews, 74, 250–72.

553   Klausen, J., Troeber, S.P., Haderlein, S.B., and Schwarzenback, R.P. (1995) Reduction of

554   Substituted Nitrobenzenes by Fe(II) in Aqueous Mineral Suspensions. Environmental

555   Science and Technology, 29, 2396–2404.

556   Lagroix, F., Banerjee, S.K., and Jackson, M.J. (2016) Geological Occurrences and Relevance of

557   Iron Oxides. In Iron Oxides: From Nature to Applications p. 7–30., Wiley - VCH,

558   Weinheim, Germany.

559   Larese-Casanova, P., and Scherer, M.M. (2007) Fe(II) sorption on hematite: New insights based

560   on spectroscopic measurements. Environmental Science and Technology, 41, 471–477.

561   Larese-Casanova, P., Kappler, A., and Haderlein, S.B. (2012) Heterogeneous oxidation of Fe(II)

562   on iron oxides in aqueous systems: Identification and controls of Fe(III) product formation.

563   Geochimica et Cosmochimica Acta, 91, 171–186.

564   Maqueda, C., Undabeytia, T., Villaverde, J., and Morillo, E. (2017) Behaviour of glyphosate in a

565   reservoir and the surrounding agricultural soils. Science of the Total Environment, 593–594,

566   787–795.

567   Navrotsky, A., Mazeina, L., and Majzlan, J. (2008) Size-Driven Structural and Thermodynamic

568   Complexity in Iron Oxides. Science, 319, 1635–1638.

569   Pauling, L., and Hendricks, S.B. (1925) The Crystal Structures of Hematite and Corundum.

570   Journal of the American Ceramic Society, 47, 781-790.

571   Pecher, K., Haderlein, S.B., and Schwarzenbach, R.P. (2002) Reduction of polyhalogenated

572   methanes by surface-bound Fe(II) in aqueous suspensions of iron oxides. Environmental

573   Science and Technology, 36, 1734–1741.

574   Pedersen, H.D., Postma, D., Jakobsen, R., and Larsen, O. (2005) Fast transformation of iron

575   oxyhydroxides by the catalytic action of aqueous Fe(II). Geochimica et Cosmochimica

576   Acta, 69, 3967–3977.

577   Penn, R.L., Erbs, J.J., and Gulliver, D.M. (2006) Controlled Growth of alpha-FeOOH Nanorods

578   by Exploiting-Oriented Aggregation. Journal of Crystal Growth, 293, 1–4.

579   Rosso, K.M., Smith, D.M.A., and Dupuis, M. (2003) An ab initio model of electron transport in

580   hematite (α-Fe2O3) basal planes. Journal of Chemical Physics, 118, 6455–6466.

581   Rosso, K.M., Yanina, S. V., Gorski, C.A., Larese-Casanova, P., and Scherer, M.M. (2010)

582   Connecting observations of hematite (α-Fe2O3) growth catalyzed by Fe(II). Environmental

583   Science and Technology, 44, 61–67.

584   Scherer, M.M., Richter, S., Valentine, R.L., and Alvarez, P.J.J. (2000) Chemistry and

585   Microbiology of Permeable Reactive Barriers for In Situ Groundwater Clean up. Critical

586   Reviews in Environmental Science and Technology, 30, 363–411.

587   Schwertmann, U., and Murad, E. (1983) Effect of pH on the Formation of Goethite and Hematite

588   from Ferrihydrite. Clays and Clay Minerals, 31, 277–284.

589   Stemig, A.M., Do, T.A., Yuwono, V.M., Arnold, W.A., and Penn, R.L. (2014) Goethite

590   Nanoparticle Aggregation: Effects of Buffers, Metal Ions, and 4-Chloronitrobenzene

591   Reduction. Environmental Science: Nano, 1, 478–487.

592   Strehlau, J.H., Stemig, M.S., Penn, R.L., and Arnold, W.A. (2016) Facet-Dependent Oxidative

593   Goethite Growth As a Function of Aqueous Solution Conditions. Environmental Science

594   and Technology, 50, 10406–10412.

595   Strehlau, J.H., Schultz, J.D., Vindedahl, A.M., Arnold, W.A., and Penn, R.L. (2017) Effect of

596   nonreactive kaolinite on 4-chloronitrobenzene reduction by Fe(II) in goethite–kaolinite

597   heterogeneous suspensions. Environmental Science: Nano, 4, 325–334.

598   Szytula, A., Bureqicz, A., Dimitrijevic, Z., Krasnicki, S., Rzany, H., Todorovic, J., Wanic, A.,

599   and Wolski, W. (1968) Neutron Diffraction Studies of α-FeOOH. Physica Status Solidi, 26,

600   429-434.

601   Tanwar, K.S., Petitto, S.C., Ghose, S.K., Eng, P.J., and Trainor, T.P. (2008) Structural study of

602   Fe(II) adsorption on hematite(1-102). Geochimica et Cosmochimica Acta, 72, 3311–3325.

603   Valencia-Avellan, M., Slack, R., Stockdale, A., and Mortimer, R.J.G. (2017) Understanding the

604   mobilisation of metal pollution associated with historical mining in a carboniferous upland

605   catchment. Environmental Science: Processes Impacts, 19, 1061–1074.

606   Vindedahl, A.M., Arnold, W.A., and Penn, R.L. (2015) Impact of pahokee peat humic acid and

607   buffer identity on goethite aggregation and reactivity. Environmental Science: Nano, 2,

608   509–517.

609   Wehrli, B., Sulzberger, B., and Stumm, W. (1989) Redox processes catalyzed by hydrous oxide

610   surfaces. Chemical Geology, 78, 167–179.

611   West, A.R. (2014) Solid State Chemistry and its Applications, p. 248-258. Wiley, U.K.

612   Williams, A.G.B., and Scherer, M.M. (2004) Spectroscopic Evidence for Fe(II)-Fe(III) Electron

613   Fransfer at the Iron Oxide-Water Interface. Environmental Science and Technology, 38,

614   4782–4790.

615   Yang, H., Lu, R., Downs, R.T., and Costin, G. (2006) Goethite, α-FeO(OH), from single-crystal

616   data. Acta Crystallographica Section E: Structure Reports Online, 62.

617

618

619

27

620

621
622
623
624
625
626
627 **Table 1.** Summary of reaction conditions in suspensions of rhombohedral (R) and

628 equidimensional (E) hematite, with final goethite mass percent in post-reaction material

629 determined using the XRD calibration curve (Fig. S11). Abbreviations are surface area loading

630 (S.A. Loading), stoichiometric maximum of goethite (Gth Max.), observed goethite (Gth Obs.),

631 and standard deviation of the observed goethite mass% across replicate reactors and/or replicate

632 XRD scans (St Dev). Dash symbol (-) indicates not detected.

633

Set	ID	Mass Loading (mg/mL)	S.A. Loading (m^2/mL)	# Exposures, ([4-ClNB] µM)	pH	Fe(II) (mM)	Gth Max (mass%)	Gth Obs (mass%)	St. Dev. (mass%)
Ref.	R-r	0.100	0.0072	1, (100)	7.0	1.00	35	15.4	2.2
a	R-a1	0.200	0.0145	1, (100)	7.0	1.00	21	trace	
	R-a2	0.275	0.0199	1, (100)	7.0	1.00	16	-	
	R-a3	0.300	0.0217	1, (100)	7.0	1.00	15	-	
	R-a4	0.325	0.0235	1, (100)	7.0	1.00	14	-	
b	R-b1	0.300	0.0217	3, (100)	7.0	1.00	35	9.3	0.8
	R-b2	0.600	0.0434	6, (100)	7.0	1.00	35	1.9	0.6
c	R-c1	0.100	0.0072	1, (100)	6.5	1.00	35	trace	
d	R-d1	0.100	0.0072	2, (50)	7.0	0.50	35	9.5	1.2
	R-d2	0.100	0.0072	4, (25)	7.0	0.25	35	6.2	0.8
e	E-e1	0.383	0.0072	1, (100)	7.0	1.00	12	5.2	0.7
	E-e2	1.149	0.0217	3, (100)	7.0	1.00	12	1.9	0.4

634

635 **Table 2.** Areal density of contiguous singly-coordinated hydroxyl group pairs across four Miller
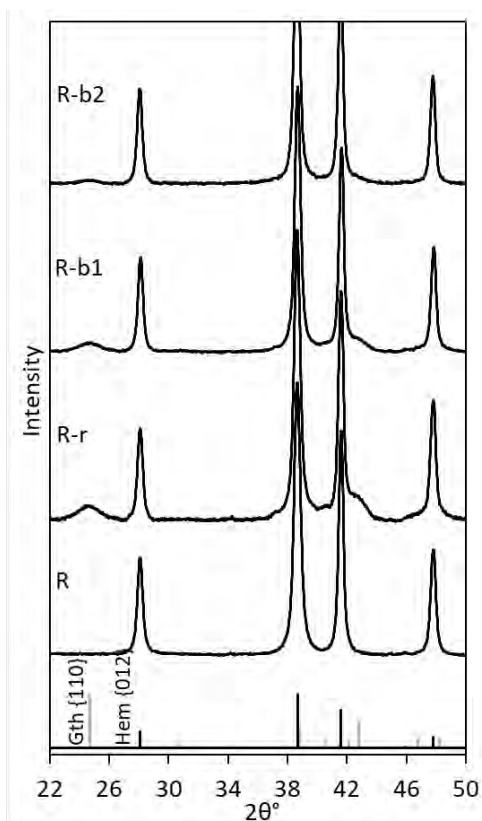
636 indices of hematite and goethite with corresponding lattice spacing (Barron and Torrent 1996).

637

28

Miller Index	d-Spacing (Å)	Areal Density of Contiguous Singly-Coordinated Hydroxyl Groups (pairs per nm^2)
Hem {113}	2.21	2.1
Hem {110}	2.52	2.5
Hem {100}	4.36	2.9
Hem {012}	3.68	3.7
Gth {110}	4.18	1.5
Gth {100}	4.61	1.7
Gth {010}	9.96	3.6
Gth {021}	2.58	4.1

638

639

640

**Figure 1.** X-ray diffraction patterns for a recurrent-exposure set using rhombohedral hematite.

Refer to Table 1 for sample identifiers. The diffraction pattern of the pure rhombohedral stock

suspension is denoted R. Origin (OriginLab Corporation 2016, version 93E) was used to

644   subtract the baseline, smooth the data using 3 average-adjacent data points, and normalize each

645   pattern to the hematite {012} peak intensity. See Fig. S9 for the raw pattern of sample R-r.

646

647

648
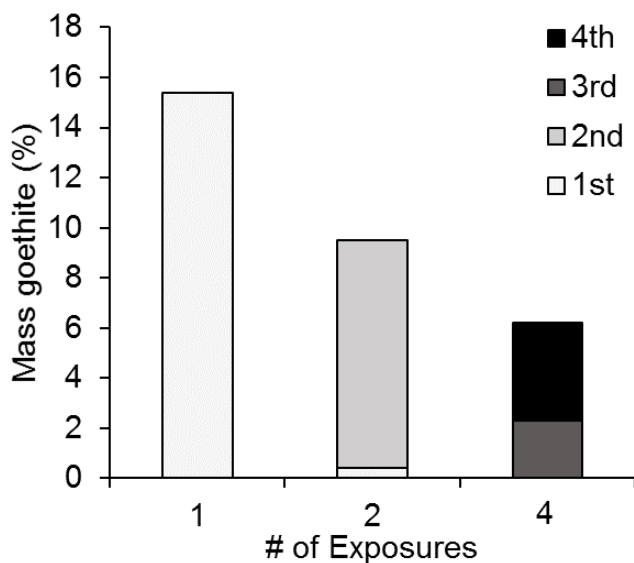
649

650   **Figure 2.** Mass% of goethite present after each exposure of 4-ClNB in the single-exposure

651   reference and in the recurrent-exposure reactors of Set d, where the Fe(II) and 4-ClNB

652   concentrations were halved and quartered as compared to the reference condition and given two

653   and four exposures, respectively.

654
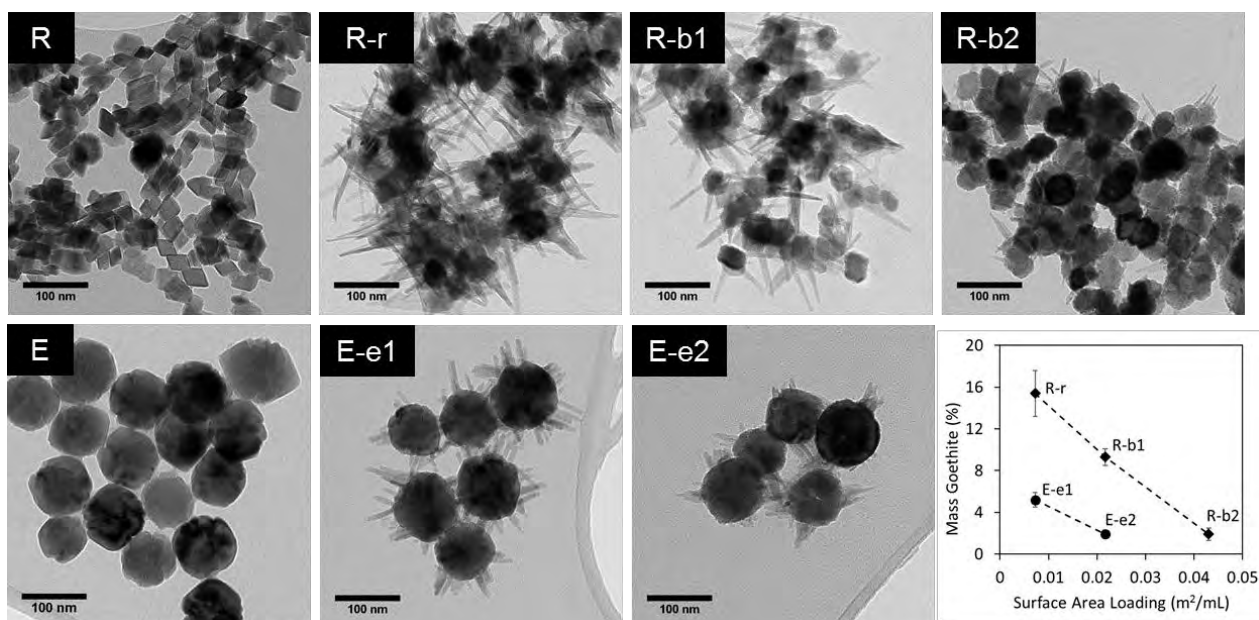
Figure 3. TEM images of solid materials collected from batch reactors prepared with the synthetic rhombohedral (R) and equidimensional (E) hematite particles. See Table 1 for corresponding sample identifiers, reactor conditions, and mineral phase quantification. The plot shows observed goethite mass percent vs reactor surface area loading for each post-reaction material shown in the given TEM images. Dashed lines are used to guide the eye for each reaction set. Error bars represent the standard deviation of replicate reactors and/or XRD scans.
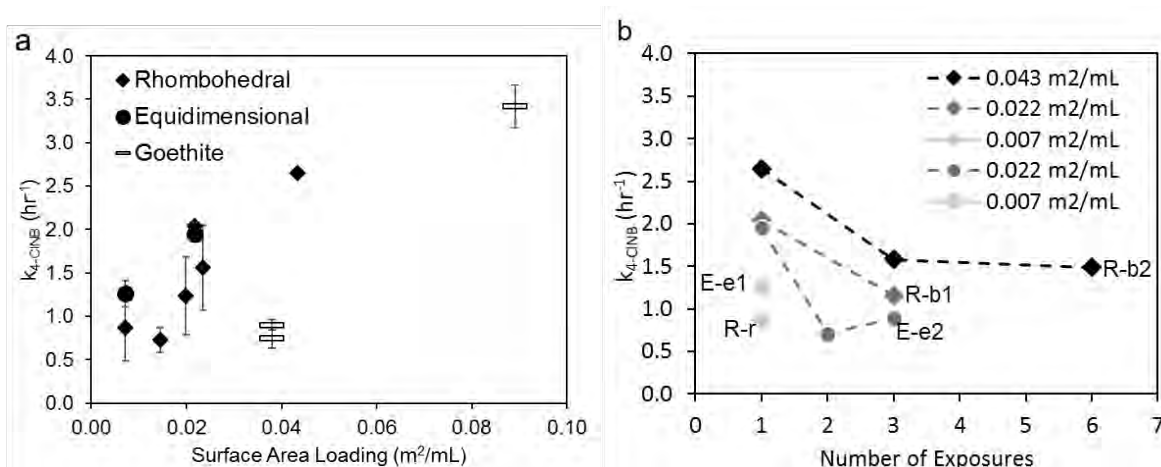
Figure 4. Reaction rate data for (a) single- and (b) recurrent-exposures of 4-ClNB on hematite nanoparticles of rhombohedral and equidimensional morphologies. Error bars represent the 95

31

666   % confidence interval. Dashed lines in (b) are intended to guide the eye among exposure

667   sequences. Goethite rate data points are literature values with their reported 95 % confidence

668   intervals (Chun et al. 2006; Stemig et al. 2014; Vindedahl et al. 2015).

669

670

671
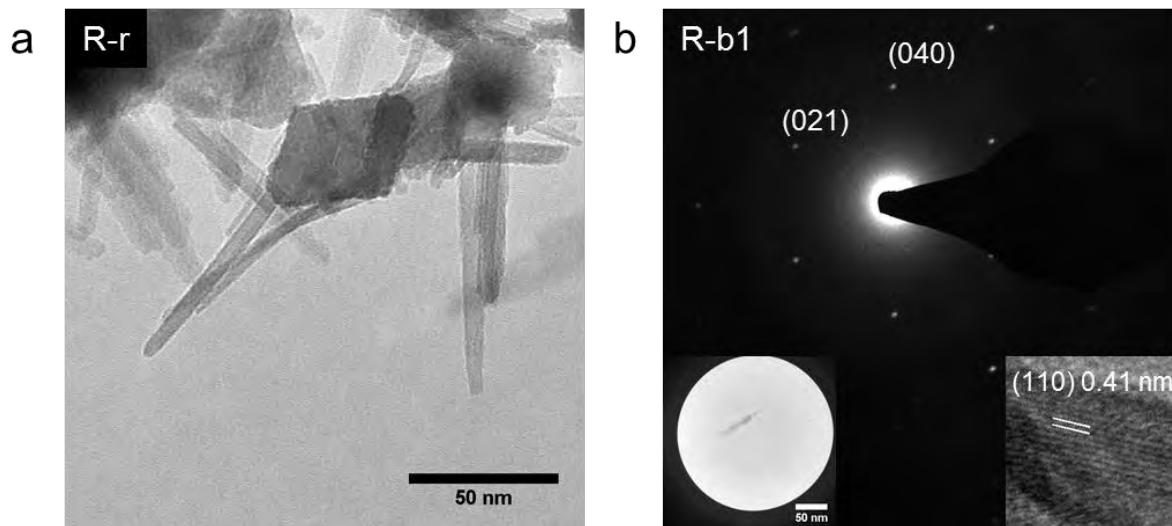
672
673   **Figure 5.** TEM images showing (a) the kite-shaped particle morphology post-reaction and (b)

674   electron diffraction pattern and crystal lattice measurements (right inset), which indicate the

675   presence of goethite. The left inset shows a selected rod aggregate that produced the diffraction

676   pattern viewed down the [100] zone axis. Additional electron diffraction patterns and crystal

677   lattice measurements for other reactors are shown in Fig. S8.

678

679

32
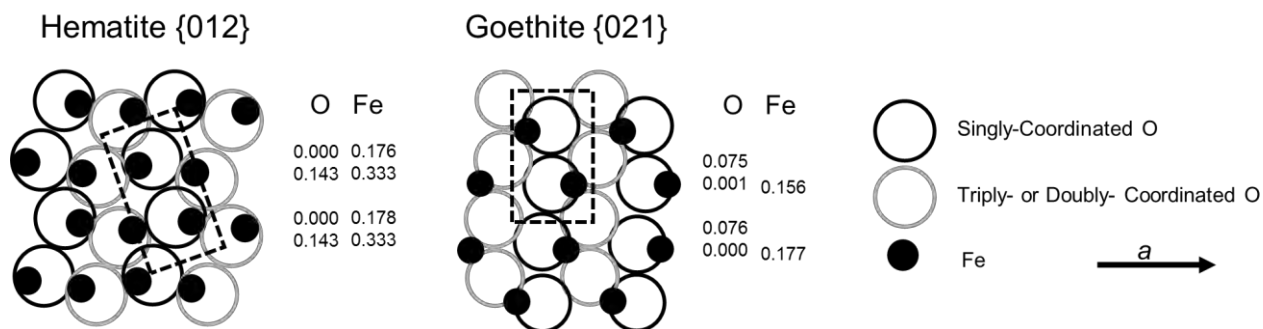
680

681 **Figure 6.** Schematic of the hematite {012} and goethite {021} surface Fe and O positions. Gray

682 open circles indicate singly-coordinated hydroxyl groups. Black open circles indicate triply-

683 coordinated hydroxyl groups on hematite and doubly-coordinated hydroxyl groups on goethite.

684 Dotted-line boxes contain a pair of contiguous singly-coordinated hydroxyl groups. Columns of

685 numbers to the right of each schematic are the distances (in nm) of the respective atoms (O or

686 Fe) below the surface as defined by the positions of the oxygen anions labeled with 0.000. Both

687 surfaces are positioned according to the *a* unit cell axis. Adapted from Barron and Torrent,

688 1996.