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Revision 1

Uptake and Release of Arsenic and Antimony in Alunite-Jarosite and Beudantite Group

Minerals

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14 **ABSTRACT**

15 Arsenic and antimony are highly toxic to humans, animals and plants. Incorporation in alunite, jarosite
16 and beudantite group minerals can immobilize these elements and restrict their bioavailability in acidic,
17 oxidizing environments. This paper reviews research on the magnitude and mechanisms of
18 incorporation of As and Sb in, and release from, alunite, jarosite and beudantite group minerals in
19 mostly abiotic systems. Arsenate-for-sulfate substitution is observed for all three mineral groups, with
20 the magnitude of incorporation being beudantite (3-8.5 wt. % As) > alunite (3.6 wt. % As) >
21 natroalunite (2.8 wt. %) > jarosite (1.6 wt. % As) > natroalunite (1.5 wt. % As) > hydroniumalunite
22 (0.034 wt. % As). Arsenate substitution is limited by the charge differences between sulfate (-2) and
23 arsenate (-3), deficiencies in B-cations in octahedral sites and for hydroniumalunite, difficulty in
24 substituting protonated H₂O-for-OH⁻ groups. Substitution of arsenate causes increases in the *c*-axis for
25 alunite and natroalunite, and in the *c*- and *a*-axes for jarosite. The degree of uptake is dependent on, but
26 limited by, the AsO₄/TO₄ ratio. Aerobic and abiotic As release from alunite and natroalunite is limited,
27 especially between pH 5 and 8. Release of As is also very limited in As-bearing jarosite, natrojarosite
28 and ammoniumjarosite at pH 8 due to formation of secondary maghemite, goethite, hematite and Fe
29 arsenates that resorb the liberated As. Abiotic reductive dissolution of As-bearing jarosite at pH 4, 5.5
30 and 7 is likewise restricted by the formation of secondary green rust sulfate, goethite and lepidocrocite
31 that take up the As. Similar processes have been observed for the aerobic dissolution of Pb-As-jarosite
32 (beudantite analogue), with secondary Fe oxyhydroxides resorbing the released As at pH 8. Higher
33 amounts of As are released, however, during microbial-driven jarosite dissolution. Natural jarosite has
34 been found to contain up to 5.9 wt. % Sb⁵⁺ substituting for Fe³⁺ in the B-site of the mineral structure.
35 Sb(V) is not released from jarosite at pH 4 during abiotic reductive dissolution, but at pH 5.5 and 7, up
36 to 75% of the mobilized Sb can be structurally incorporated into secondary green rust sulfate,
37 lepidocrocite or goethite. Further research is needed on the co-incorporation of As, Sb and other ions

38 in, and the uptake and release of Sb from, alunite, jarosite and beudantite group minerals, the influence
39 of microbes on these processes and the long-term (>1 year) stability of these minerals.

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INTRODUCTION

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43 Arsenic (As) and antimony (Sb) are two of the most toxic elements to humans and many biota
44 (Sundar and Chakravarty 2010; Feng et al. 2013; Abdul et al. 2015). The risks posed by exposure to As
45 and Sb can be mitigated by the precipitation of minerals and especially, those that are relatively
46 insoluble and have low bioaccessibility. Alunite, jarosite and beudantite-group minerals form in
47 oxidizing, low-pH conditions in a wide variety of environments, including mine wastes (Hudson-
48 Edwards et al. 1999; Kocourková et al., 2011; Nieto et al. 2003), acid sulfate soils (Nordstrom 1982;
49 Welch et al. 2007), saline lakes (Alpers et al. 1992) and sulfuric acid caves (D'Angeli et al. 2018). A
50 significant body of work has been carried out to understand the capacity and mechanisms of uptake of
51 As and Sb in, and release from, alunite, jarosite and beudantite-group minerals formed in these
52 environments. The aim of this paper is to synthesize and review research, and to outline research gaps
53 that should be filled by future work.

54

MINERALS OF THE ALUNITE, JAROSITE AND BEUDANTITE GROUPS

55

56 Minerals of the alunite supergroup have a general formula of $AB_3(TO_4)_2(OH)_6$. In the formula,
57 A represents cations with a coordination number greater than or equal to 9, B represents cations in a
58 slightly distorted octahedral coordination, and T represents cations with tetrahedral coordination
59 (Kubisz 1964, 1970; Jambor 1999; Hawthorne et al. 2000; Fig. 1). Within the supergroup, the alunite
60 group minerals contain more Al^{3+} than Fe^{3+} , and jarosite group minerals contain more Fe^{3+} than Al^{3+} ,
61 in the B-site (Table 1). The A-site can be filled by univalent or divalent cations such as K^+ , Na^+ , H_3O^+ ,
62 Ca^{2+} , Sr^{2+} , Ba^{2+} and Pb^{2+} , but the most common substitutions are K^+ (giving alunite and jarosite), Na^+
63 (giving natroalunite and natrojarosite) and H_3O^+ (giving hydroniumalunite and hydroniumjarosite).
64 When divalent cations substitute for monovalent cations, charge balance is maintained in the alunite-
65 jarosite structure by creating A-site vacancies with replacement of two monovalent cations by one

66 divalent cation (e.g., plumbojarosite), the incorporation of divalent cations in the B-site (e.g.,
67 beaverite), or the incorporation of trivalent anions in normally divalent anion sites (e.g., beudantite).
68 The T site of alunite supergroup minerals is normally filled by S^{6+} in the sulfate anion SO_4^{2-} , but can
69 also be filled by As^{5+} in AsO_4^{3-} , P^{5+} in PO_4^{3-} and Si^{4+} in SiO_2 . The beudandite group contains both
70 SO_4^{2-} and AsO_4^{3-} in the T site, Al^{3+} , Fe^{3+} or Ga^{3+} in the B site, and Ba^{2+} , Pb^{2+} or Sr^{2+} in the B site
71 (Table 1). Solid solutions have been reported to exist between jarosite, plumbojarosite and beudantite
72 (Sánchez et al. 1996).

73 Minerals of the alunite supergroup have $R\bar{3}M$ symmetry, and are formed by (SO_4) tetrahedral -
74 (MO_6) octahedral sheets with $[M_3(OH)_6(SO_4)_2]$ compositions, which are linked by interstitial cations
75 and hydrogen bonds (Hawthorne et al. 2000; Fig. 1). Each of the octahedra are slightly distorted, and
76 have four bridging hydroxyl groups lying in a plane with sulfate oxygens at the apices. Three of the
77 SO_4 tetrahedra are coordinated to Fe octahedra, causing lowering of the SO_4 symmetry from T_d to C_{3v} -
78 (Jambor 1999; Becker and Gasharova 2001). The mineral structures often contain A- and B-site
79 vacancies (Dutrizac and Jambor 2000), and 'excess', non-OH water, which is assumed to be in the
80 form of hydronium (H_3O^+) that substitutes for K^+ on the A-site (e.g. Ripmeester et al. 1986; Gale et al.
81 2010).

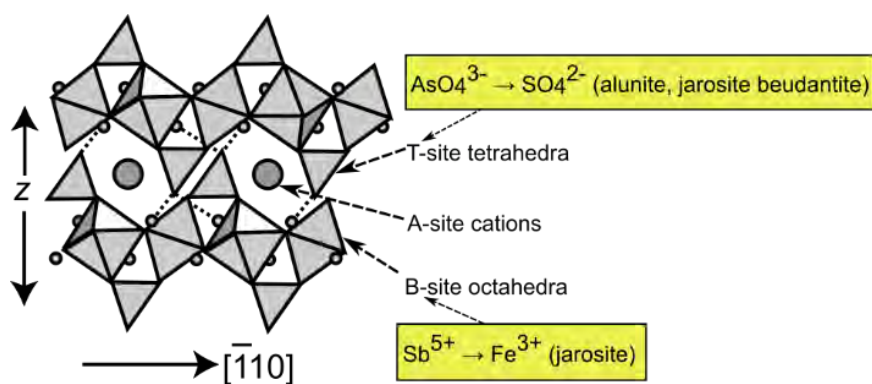


FIGURE 1. Structure of minerals of the alunite-jarosite and beudantite groups, showing sites where AsO_4^{3-} and Sb^{5+} substitute in the structure (see text for details).

88 **TABLE 1.** Alunite, jarosite and beudantite family minerals reported to contain As and Sb as major and
89 trace constituents.

Mineral	Formula
<i>Beudantite group minerals with As as a major component</i>	
Beudantite	$\text{PbFe}_3(\text{AsO}_4)(\text{SO}_4)(\text{OH})_6$
Gallobeutantite	$\text{PbGa}_3(\text{AsO}_4)(\text{SO}_4)(\text{OH})_6$
Hidalgoite	$\text{PbAl}_3(\text{AsO}_4)(\text{SO}_4)(\text{OH})_6$
Kemmlitzite	$\text{SrAl}_3(\text{AsO}_4)(\text{SO}_4)(\text{OH})_6$
<i>Alunite-jarosite group minerals with As as a trace component</i>	
Alunite	$\text{KAl}_3(\text{SO}_4)_2(\text{OH})_6$
Ammoniojarosite	$\text{NH}_4\text{Al}_3(\text{SO}_4)_2(\text{OH})_6$
Natroalunite	$\text{NaAl}_3(\text{SO}_4)_2(\text{OH})_6$
Hydroniumalunite	$\text{H}_3\text{OAl}_3(\text{SO}_4)_2(\text{OH})_6$
Jarosite	$\text{KFe}_3(\text{SO}_4)_2(\text{OH})_6$
Natrojarosite	$\text{NaFe}_3(\text{SO}_4)_2(\text{OH})_6$
Plumbojarosite	$\text{K}(\text{Pb,Fe})_3(\text{SO}_4)_2(\text{OH})_6$
<i>Jarosite and Beudantite group minerals with Sb as a trace component</i>	
Jarosite	$\text{KFe}_3(\text{SO}_4)_2(\text{OH})_6$
Beudantite	$\text{PbFe}_3(\text{AsO}_4)(\text{SO}_4)(\text{OH})_6$

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92 **UPTAKE OF ARSENIC IN ALUNITE-JAROSITE AND BEUDANTITE GROUP MINERALS**

93 **Alunite minerals**

94 To the author's knowledge, no papers have reported the occurrence of natural arsenical
95 members of the alunite family. Arsenate-for-substitution in alunite, natroalunite and hydroniumalunite
96 has been investigated experimentally (Sunyer and Viñals 2011a, b; Sunyer et al. 2013; Luo et al. 2015).
97 Hydroniumalunite takes only up 1 mole % ($\text{AsO}_4/\text{SO}_4+\text{AsO}_4$) (0.034 wt. % As), possibly due to the
98 difficulty in substituting protonated H_2O -for- OH^- groups in the structure due to blocking by H-bridges
99 of the H_3O (Sunyer et al. 2013). Arsenate substitution is much higher in alunite and natroalunite (up to
100 3.61 and 2.80 wt. % As, respectively) formed at pH 1 and 2.8-2.9, respectively, and 200 °C (Sunyer et
101 al. 2013; Sunyer and Viñals 2011a). Luo et al. (2015) obtained a slightly lower maximum incorporation
102 of arsenate in natroalunite, giving an approximate molar ratio of 1:1 ($\text{AsO}_4/\text{SO}_4+\text{AsO}_4$) % at pH 3.00
103 and 200 °C.

104 The degree of uptake of As in alunite and natroalunite is directly related to the AsO_4/TO_4 ratio
105 of the initial solution. Arsenical natroalunite precipitation is favoured at $(\text{AsO}_4/\text{TO}_4)_{\text{aq}} < 0.2$ at 200 °C.
106 At higher $(\text{AsO}_4/\text{TO}_4)_{\text{aq}}$, other arsenate phases such as alarsite (AlAsO_4), mansfieldite ($\text{AlAsO}_4 \cdot 2\text{H}_2\text{O}$)
107 and natropharmacoalumite ($\text{NaAl}_4(\text{OH})_4(\text{AsO}_4)_3 \cdot 4\text{H}_2\text{O}$) form (Sunyer and Viñals 2011a). At 200 °C,
108 the ratio of $(\text{AsO}_4/\text{TO}_4)$ in the precipitated natroalunite was roughly equivalent to $0.5(\text{AsO}_4/\text{TO}_4)_{\text{aq}}$.
109 Precipitation rates also increased with increasing temperature and AsO_4^{3-} concentrations. Experiments
110 in which gypsum was introduced at the outset resulted in an $(\text{Ca}/\text{Ca}+\text{Na})$ molar ratio of 4-6% Ca-for-
111 Na substitution and arsenate-for-sulfate substitution.

112 The degree of arsenate uptake in alunite was also dependent on $(\text{AsO}_4/\text{TO}_4)_{\text{aq}}$, with the degree
113 of substitution increasing as $0.5(\text{AsO}_4/\text{TO}_4)_{\text{alunite}} \approx 0.5(\text{AsO}_4/\text{TO}_4)_{\text{aq}}$ (Sunyer et al. 2013). Arsenical
114 alunite was the only phase to precipitate at $(\text{AsO}_4/\text{TO}_4)_{\text{aq}} < 0.26$ at 200 °C; above this ratio mansfieldite
115 co-precipitated.

116 Viñals et al. (2010) precipitated a complex natroalunite ($\sim(\text{Na,Ca})(\text{Al,Fe})_3((\text{S,As,P})\text{O}_4)_2(\text{OH})_6$ at
117 $(\text{Al}/\text{As})_{\text{aq}} = 4.5$) by reacting H_2SO_4 -leached and ozoned calcium arsenate wastes produced from copper
118 smelting at 180-200 °C with sodium and aluminum sulfate reagents. At pH 2.8-2.9, the As substitution
119 in the tetrahedral site was 7-8 mole % $(\text{AsO}_4/\text{SO}_4+\text{AsO}_4)$. The degree of substitution increased to a
120 maximum molar ratio of 14% with decreasing $(\text{Al}/\text{As})_{\text{aq}}$, but mansfieldite also precipitated. Prior
121 gypsum removal and As concentrations of 3.5-7.0 g/L were shown to not affect As incorporation or the
122 nature of the As-bearing phases formed.

123 The effect of arsenate substitution on the alunite and natroalunite unit cell was investigated by
124 Rietveld refinement. Sunyer and Viñals (2011b) showed that the *c*-parameter slightly increased with
125 arsenate for sulfate substitution due to differences between As-O and S-O distances in the tetrahedral
126 site. In contrast, the *a*-parameter was unaffected by arsenate incorporation. The same findings for
127 alunite were reported by Sunyer et al. (2013).

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129

130 **Jarosite minerals**

131 Naturally-occurring arsenian jarosite has been observed in mine wastes (Acero et al., 2006;
132 Hudson-Edwards et al. 1999, 2005; Kocourková et al., 2011; Savage et al. 2000; Filippi et al. 2015),
133 with concentrations of As up to 2000 pm reported (Savage et al., 2000). Most insights into As
134 incorporation mechanisms into jarosite, however, have been studied experimentally (Paktunc and
135 Jambor 2003; Savage et al. 2005; Karimian et al. 2017). This work has consistently shown that AsO_4^{3-}
136 occupies tetrahedral SO_4^{2-} sites within the jarosite structure. X-ray diffraction and XANES analysis
137 have shown that at least 9.9 wt. % AsO_4^{3-} (1.6 wt. % As) can be structurally incorporated into synthetic
138 jarosite (Paktunc and Dutrizac 2003), with others demonstrating that up to 30% replacement of SO_4^{2-}
139 by AsO_4^{3-} can occur (Savage et al. 2005; Paktunc et al. 2003; Kendall et al. 2013). The limiting factors
140 are proposed to be charge imbalances between sulfate and arsenate, and deficiencies of Fe in octahedral
141 sites (Savage et al., 2005). EXAFS data modelling suggests As-O interatomic distances of 1.68 Å and
142 coordination numbers of 4.6+/- 1.7 to 5.4+/-1.7, confirming the tetrahedral arrangement of O atoms
143 around central As atoms (Paktunc and Dutrizac, 2003). Incorporation of arsenate for sulfate causes
144 little change in the *a*-axis, but lengthening of the *c*-axis up to 0.174 Å (Paktunc and Dutrizac 2003;
145 Savage et al. 2005). Jarosite morphology also becomes more anhedral with increasing incorporation of
146 As (Savage et al. 2005).

147 Co-incorporation of Pb^{2+} results in a larger proportion of As^{5+} being incorporated in the jarosite
148 structure (33% of the tetrahedral sites) compared to when Pb^{2+} is not incorporated (21%; Aguilar-
149 Carrillo et al. 2018). This was proposed to be due to changes in unit cell dimensions to balance the
150 distortion caused by the substitution of arsenate for sulfate in jarosite. The concentrations of As and Pb
151 in the initial experimental solutions limited As and Pb incorporation. At $\text{As/Pb} < 1$, Pb-As jarosites

152 (beudantite analogues) formed, but at As/Pb > 1, anglesite and poorly crystalline ferric arsenate phases
153 formed along with the As-Pb jarosite.

154 Pure synthetic natrojarosites have been shown to have very low amounts of arsenate for sulfate
155 substitution (1.5 wt. %; Dutrizac and Jambor 1987). By contrast, synthetic ammoniumjarosite has been
156 shown to take up to 4.1 wt. % arsenate at a pH of 1.2-1.8 (Flores et al. 2016). The mineral, with a
157 formula of $(\text{NH}_4)\text{Fe}_{2.45}[(\text{SO}_4)_{1.80}(\text{AsO}_4)_{0.20}][(\text{OH})_{4.15}(\text{H}_2\text{O})_{1.85}]$, was shown to remain in residual solids
158 above 700 °C.

159 Using extended X-ray absorption fine structure (EXAFS) analysis, Gräfe et al. (2008) showed
160 that As^{5+} sorbed to the jarosite surface by forming bidentate-binuclear surface complexes, as it did with
161 goethite. The edge-sharing coordination number was larger for jarosite than for goethite (0.9 Fe atoms
162 at 2.87 Å and 0.3 Fe neighbours at 2.87 Å, respectively). By contrast, the radial distance to the next
163 nearest Fe atom was lower in jarosite (3.22 Å) than in goethite (3.28 Å). These results were attributed
164 to the replacement of SO_4 tetrahedra by AsO_4 tetrahedra. In the presence of Cu^{2+} , As^{5+} was shown to be
165 coordinated by multiple Cu atoms over multiple radial distances, which was modelled to be due to the
166 formation of a euchroite-like $[\text{Cu}_2(\text{AsO}_4)(\text{OH}) \cdot 3\text{H}_2\text{O}]$ complex on the jarosite surface.

167 Several studies have considered the role of microbes in the formation of arsenical jarosite. Egal
168 et al. (2009), for example, demonstrated that As-bearing jarosite formed in both biotic and abiotic
169 experiments, suggesting that bacterial iron oxidation was not involved in the process. However, aging
170 and of As-rich schwertmannite produced by microbial oxidation of Fe^{2+} in sulfate solutions yielded As-
171 bearing jarosite at lower pH values, and this might be considered to be indirectly biotically-formed.
172 The jarosite does not take up the As^{3+} released by the schwertmannite.

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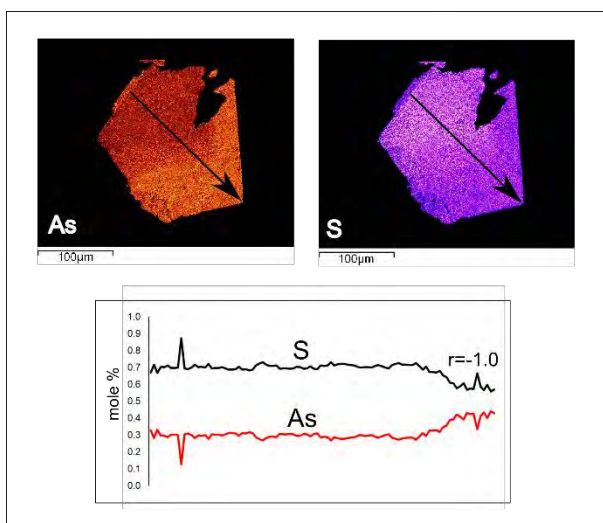
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176 **Beudantite minerals**

177 Most of the research on arsenic incorporation in beudantite group minerals (Table 1) has been
178 done on beudantite itself. Beudantite has been shown to host significant amounts (3-8.5 wt. %) of As in
179 supergene zones (Szymanski, 1988), oxidized massive sulfides and related mine wastes (Foster et al.
180 1998; Roca et al. 1999; Nieto et al. 2003; Roussel et al. 2000; Gieré et al., 2003; Courtin-Nomade et al.
181 2016) and metallurgical products (Bigam and Nordstrom 2000). Thus, it has been suggested to
182 immobilize these elements and limit their uptake by plants (e.g., Zheng et al. 2003). In high-sulfide
183 waste at the Berikul Au mine, Kemerovo region, Russia, for example, jarosite-beudantite solid solution
184 minerals have been reported to contain up to 8.5 wt. % As (Gieré et al. 2003). A beudantite from the
185 Tsumeb deposit, Namibia (Natural History Museum collection, London, UK, number BM.1987),
186 containing between 3.4 and 6.4 wt. % As, exhibits zoning in As and S, with molar proportions of these
187 elements exactly mirroring each other ($r=-1.0$; Fig. 2).

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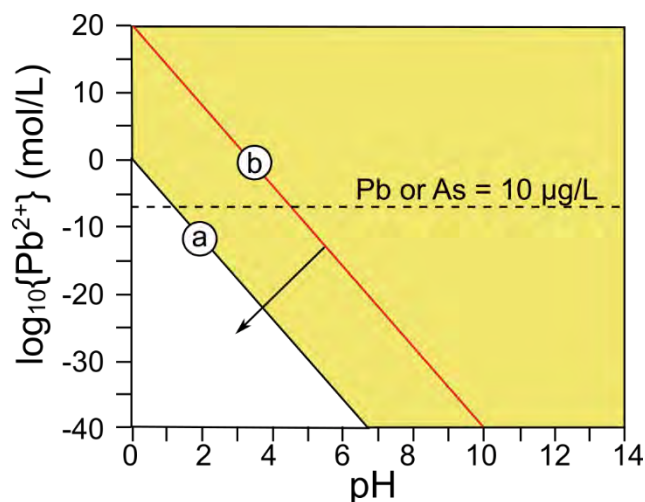
190 **FIGURE 2.** Electron microprobe X-ray chemical maps showing zoning of As and S. Black arrow
191 indicates position of line scan of analytical points, for which data are shown in the X-Y plot. Data from
192 K. Hudson-Edwards for sample number BM.1987 from the collections of the Natural History Museum,
193 London, UK. The sample was collected from the 36th level, west 12, of the Tsumeb mine, Grootfontein,
194 Namibia.

195

196 Other beudantite-group minerals have also been shown to contain significant amounts of As. A
197 sample of hidalgoite from the Tsumeb mine, Namibia, was shown to contain between 7.4 and 10.0 wt.
198 % As (Cooper and Hawthorne 2012). Arsenic-rich kemmlitzite with a T-site formula of
199 $[(AsO_4)_{0.98}(PO_4)_{0.42}(SO_4)_{0.39}(SiO_4)_{0.18}]$ has also been analysed (Jambor 1999). Similarly high amounts
200 of As were found in gallobaudantite from Tsumeb, Namibia, with a formula of
201 $Pb_{1.04}(Ga_{1.39}Fe_{0.82}Al_{0.62}Zn_{0.10})_{3.03}[(AsO_4)_{1.14}(SO_4)_{0.86}]_{2.00}(OH)_{5.94}$ (Jambor et al. 1996).

202 Symanski (1988) showed that the T-site in beudantite, which was filled by arsenate and sulfate,
203 was disordered, and suggested this was due to that the excess negative charge in arsenate compared to
204 sulfate. He further proposed that this disorder was balanced by substitution of hydronium for Pb or by
205 protonation of hydroxyl groups to water.

206 Forray et al. (2014) derived thermochemical data ($\Delta H^\circ_f = -3691.2 \pm 8.6$ kJ/mol; $\Delta G^\circ_f = -$
207 3164.78 ± 9.1 kJ/mol) and a log K_{sp} value of -13.94 ± 13.94 for the Pb-As jarosite synthesized by
208 Smith et al. (2006). The Pb-As-jarosite had a structure matching that of beudantite (XRD file PDF 19-
209 0689), and composition of $(H_3O)_{0.68}Pb_{0.32}Fe_{2.86}(SO_4)_{1.69}(AsO_4)_{0.31}(OH)_{5.59}(H_2O)_{0.41}$; thus, it did not
210 contain enough Pb or As to be considered true beudantite) The data were used to show that the
211 formation of Pb-As jarosite can decrease aqueous As and Pb concentrations to meet WHO drinking
212 water standards of 10 $\mu\text{g/L}$ (WHO, 1996, 1998; Fig. 3).



213

214 **FIGURE 3.** Stability of Pb–As jarosite as function of pH and lead activity. Point (a) shows the stability
215 boundary for Pb-As jarosite at $a(\text{Fe}^{3+}) = 322 \times 10^{-3} \text{ mol/L}$, $a(\text{As}^{5+}) = 1.33 \times 10^{-7} \text{ mol/L}$ and $a(\text{SO}_4^{2-}) =$
216 $1.56 \times 10^{-2} \text{ mol/L}$. Point (b) shows the stability boundary for Pb-As jarosite at very low Fe, SO_4 and As
217 activity ($10 \mu\text{g/L}$ each). Reprinted (adapted) from *Geochimica et Cosmochimica Acta*, 127, Forray et
218 al., Synthesis, characterization and thermochemistry of Pb-As-, Pb-Cu- and Pb-Zn-jarosite compounds,
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222 **RELEASE OF ARSENIC FROM ALUNITE-JAROSITE AND BEUDANTITE GROUP**

223

MINERALS

224 **Alunite minerals**

225 The suitability of alunite, natroalunite and hydronium alunite as short- and long-term stable stores of
226 arsenic has been explored by Viñals et al. (2010), Sunyer and Viñals (2011a, b), Sunyer et al. (2013)
227 and Luo et al. (2015). Sunyer et al. (2013) and Sunyer and Viñals (2011b) showed that after 24 h,
228 arsenical alunite and natroalunite, respectively, released only 0.01-0.05 mg/L As between pH 5 and 8.
229 (Fig. 4). Within 2.5 weeks, arsenical alunite dissolved to release 0.3 mg/L As between pH 4 and 5
230 (Sunyer et al. 2013). Longer-term stability tests carried out over 6 months by Viñals et al. (2010)
231 showed that natroalunite only released 0.01 mg/L As between pH 4 and 5. These As releases are
232 similar to, or lower than those of natural scorodite ($\text{FeAsO}_4 \cdot 2\text{H}_2\text{O}$; 0.4 mg/L As) and synthetic
233 scorodite (1.3 mg/L As) (Sunyer and Viñals 2011b). Thus, both arsenical alunite and especially,
234 natroalunite, could be considered as good long-term stores of As-bearing wastes with large SO_4^{2-}
235 / AsO_4^{3-} ratios.

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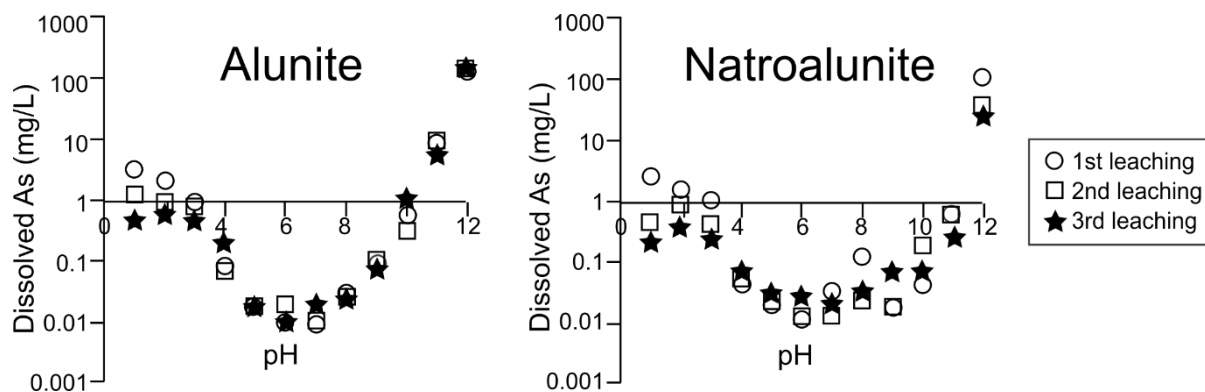
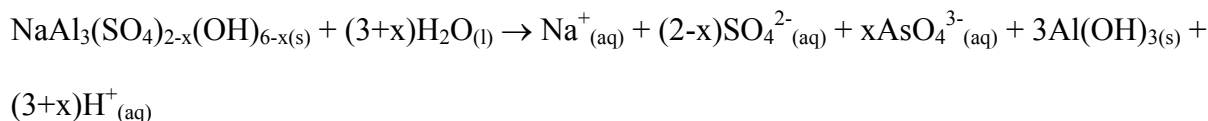


FIGURE 4. Dissolved As (mg/L) as a function of pH for alunite and natroalunite 24 h dissolution experiments. Reprinted (adapted) from Arsenic immobilization as alunite-type phases: The arsenate substitution in alunite and hydronium alunite. *Journal of Hazardous Materials*, 261, Sunyer et al., 559-569 Copyright (2013) with permission from Elsevier, and Arsenate substitution in natroalunite: A potential medium for arsenic immobilization. Part 2: Cell parameters and stability tests. *Hydrometallurgy* 109, Sunyer and Viñals, J., 106-115, Copyright (2011) with permission from Elsevier.

The dissolution of arsenical natroalunite has been shown to be congruent at pH 3 and pH 12, and incongruent at pH between 4 and 11 (Sunyer and Viñals 2011b). The mechanism proposed for the incongruent dissolution involves the formation of secondary Al hydroxides, as follows:



Jarosite minerals

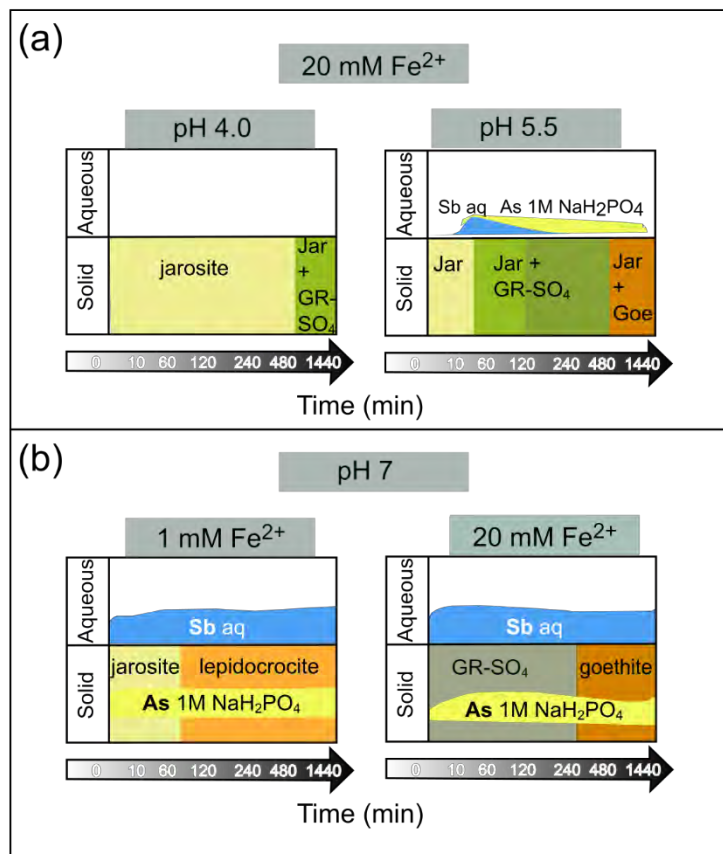
Kendall et al. (2013) investigated the dissolution of arsenojarosite in pH 2 and 8 solutions and in ultrapure water. Unlike jarosite, the dissolution of arsenojarosite was not pH-dependent. This was proposed to be due to the presence of surface arsenate-iron complexes that prevented protonation at low pH and hydroxyls at high pH. Arsenojarosite dissolution was incongruent, with maghemite, goethite and hematite forming in ultra-pure water experiments, and ferrihydrite in pH 8 Tris-buffered experiments. Short-term dissolution rates were observed to increase with increasing incorporation of arsenate. These rates were attributed to polarization of bonds within the jarosite structure due to

262 substitution of arsenate, which is larger and higher charged than sulfate. This process would make these
263 bonds more susceptible to dissolution. Preferential leaching of K^+ and SO_4^{2-} compared to Fe^{3+} (cf.,
264 Smith et al. 2006) causes enrichment of arsenate-iron complex sites on the surface. Dissolution is then
265 inhibited due to the fact that arsenate-iron complexes are more strongly bonded than sulfate-iron
266 complexes within the structure.

267 The alkaline decomposition of As-bearing jarosite minerals results in the formation of
268 secondary Fe hydroxides that take up the released As. For example, Patiño et al. (2013) showed that the
269 reaction of a synthetic sodium arsenojarsite ($Na_{0.87}(H_3O)_{0.13}Fe_{2.50}[(SO_4)_{1.95}(AsO_4)_{0.05}]$
270 $[(OH)_{4.45}(H_2O)_{1.55}]$) with CaOH and NaOH at pH 12.33-12.87 caused an $Fe(OH)_3$ with adsorbed AsO_4
271 coating to form on the original reactant. The decomposition of arsenic-bearing ammoniumjarosite using
272 NaOH at pH 12.90 and 30°C also resulted in the formation of a secondary Fe arsenate with a
273 composition of $2.45 Fe(OH)_3 \cdot 0.20 AsO_4^{3-}$ (Flores et al., 2016).

274 Karimian et al. (2017) determined the mechanisms and magnitude of As release during the
275 reductive dissolution of jarosite by Fe^{2+} at pH 4.0, 5.5 and 7. At pH 4, the jarosite did not change or
276 release As^{5+} (Fig. 5). At pH 5.5, the jarosite transformed first to green rust (after 60 min) and then to
277 goethite (after 24th), and the As released was taken up on the jarosite surface mostly by labile (c. 10%)
278 and poorly crystalline Fe^{3+} -associated phases (10-30 %). At pH 7, the jarosite transformed rapidly (<10
279 min) to green rust and then to goethite at high Fe^{2+} concentrations (10 and 20 mM), and to
280 lepidocrocite (< 60 min) at low Fe^{2+} concentrations (1 and 5 mM) (Fig. 5). Some of the original As^{5+} in
281 the jarosite was reduced to As^{3+} in the high Fe^{2+} experiments but not in the low Fe^{2+} experiments. As in
282 the pH 5.5 experiments, the low concentrations of $As_{(aq)}$ suggested that the As released during
283 dissolution was taken up by the secondary Fe^{3+} phases either by surface sorption or incorporation in
284 structural defects.

285



286

287 **FIGURE 5.** Schematic diagram showing results of reductive dissolution of As- and Sb-bearing jarosite
 288 by Fe²⁺ at pH 4, 5.5 and 7. See text for discussion. (a) Reprinted (adapted) from Antimony and arsenic
 289 partitioning during Fe²⁺-induced transformation of jarosite under acidic conditions
 290 Chemosphere, 195, Karimian et al., 515-523, Copyright (2018) with permission from Elsevier. (b)
 291 Adapted with permission from Karimian et al. (2017). Copyright 2017 American Chemical Society.
 292

293 Sulfide-promoted reductive dissolution of As⁵⁺-bearing jarosite resulted in its replacement by
 294 mackinawite and mobilization of the As (Johnson et al., 2012). Increasing S²⁻:Fe ratios and decreasing
 295 pH resulted in increased As mobility, except at high S²⁻:Fe and pH 4.0 when an orpiment-like phase
 296 (detected using transmission electron microscopy, TEM) formed. EXAFS analysis was used to show
 297 the transition from O-coordinated As⁵⁺ to S-co-ordinated As⁵⁺ in the orpiment phase.

298

T

299 Beudantite minerals

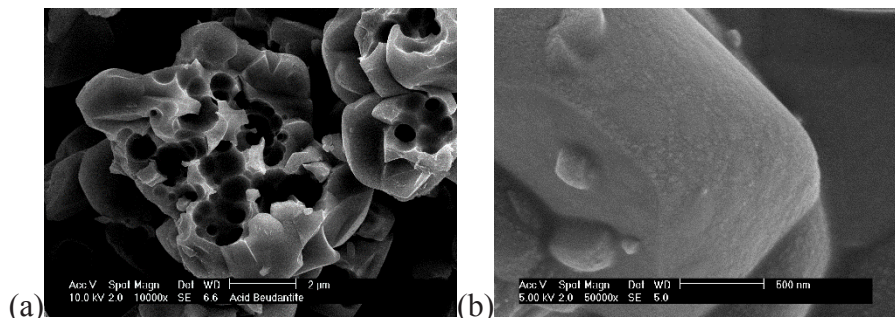
300 Batch dissolution experiments using a synthetic Pb-As-jarosite (beudantite analogue, see above)
 301 were conducted at 20 °C and pH 2 and 8, to mimic environments affected by acid rock/acid mine

302 drainage, and those remediated with slaked lime ($\text{Ca}(\text{OH})_2$), respectively (Smith et al. 2006). The
303 dissolutions were both incongruent. Dissolution at pH 2 yielded poorly crystalline solid PbSO_4 (Fig. 6)
304 and aqueous Fe , SO_4^{2-} and AsO_4^{3-} . Dissolution at pH 8, by contrast, produced $\text{Fe}(\text{OH})_3$ that resorbed
305 aqueous AsO_4^{3-} and solid PbSO_4 .

306 These results were explained by the preferential dissolution of the A- and T-sites, containing Pb and
307 SO_4^{2-} - AsO_4^{3-} , respectively, relative to the sterically remote and Fe octahedra within the T-O-T Pb-As-
308 jarosite structure. By contrast, Roca et al. (1999) suggested that the alkaline decomposition of
309 beudantite in $\text{Ca}(\text{OH})_2$ was very slow between 80 and 100 °C. They proposed a slightly different
310 mechanism for beudantite decomposition than Smith et al. (2006), involving the reaction of the
311 beudantite with OH^- and water, and production of $\text{Fe}(\text{OH})_3$, $\text{Pb}(\text{OH})_2$, an As_2O_5 gel and sulfate ions.

312 Few studies have investigated the role of microbial processes on As- or Sb- bearing alunite-
313 group minerals. One example is that of Smeaton et al. (2012), who used the Pb-As-jarosite phase
314 synthesized by Smith et al. (2006) to conduct reductive dissolution experiments using *Shewanella*
315 *putrefaciens* at circumneutral pH. The experiments resulted in immediate Fe^{3+} reduction and As^{5+}
316 reduction within 72 h. After 336 h, 20.2% and 3.0% of the total As occurred as solid and aqueous As^{3+} .
317 At this time, secondary Fe-O precipitates containing minor As and Pb formed, leaving c. 0.05 mM As^{3+}
318 and 2 mM As^{5+} in solution. Based on these results, the authors suggested that Pb-As-jarosite would not
319 be a good candidate for long-term As storage under reductive conditions. The reduction of Fe^{3+} was
320 proposed to be thermodynamically driven, while that of aqueous As^{5+} was proposed to be due to
321 microbial detoxification.

322



323

324 **FIGURE 6.** Scanning electron microscope (SEM) image of residual solids from the pH 2 dissolution of
325 Pb-As-jarosite, showing (a) extensive pitting of the surface and (b) poorly crystalline secondary PbSO₄
326 on the mineral surface. Reprinted (adapted) from Chemical Geology, 229, Smith et al., Dissolution of
327 lead- and lead-arsenic jarosites at pH 2 and 8: insights from batch experiments, 344-361, Copyright
328 (2006), with permission from Elsevier.
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330

331 UPTAKE AND RELEASE OF ANTIMONY IN JAROSITE GROUP MINERALS

332 The mechanisms and magnitudes of incorporation and release of Sb from alunite, jarosite or beudantite
333 family minerals have received much less attention than those of As. The few studies available have
334 focused on the uptake of Sb in, and release of Sb from, jarosite; none, to the author's knowledge, have
335 determined whether alunite or beudantite minerals can take up and release Sb.

336 Jamieson et al. (2005) reported minor uptake of Sb in jarosite (176 ± 159 ppm, $n=18$) forming
337 from waste waters of the Richmond Mine, Iron Mountain, California. By contrast, Courtin-Nomade et
338 al. (2012) found very high concentrations of Sb (59000 ± 21000 ppm) in secondary cryptocrystalline
339 jarosite in historic mill tailings from the French Massif Central. They attributed this to substitution of
340 Sb⁵⁺ for Fe³⁺ in the jarosite structure, analogous to that occurring in dussertite
341 (Ba(FeSb)₃(AsO₄)(OH,H₂O)₆; Kolitsch et al. 1999). Karimian et al. (2017) synthesized an Fe-deficient
342 jarosite (K:Fe:S = 1:2.5:1.9, compared to 1:3:2 for ideal jarosite) that incorporated 1.43 ppm Sb⁵⁺. They
343 also proposed that Sb⁵⁺ substituted for Fe³⁺ in the jarosite structure due to similar ionic sizes (Sb⁵⁺, 0.60
344 Å; Fe³⁺ 0.65 Å) and similar bond lengths in the Sb_vO₆ and Fe_{III}O₆ octahedra (Sb-O, 1.91 Å; Fe-O,
345 1.97-2 Å; Mitsunobu et al. 2013). Such Sb⁵⁺ substitution for Fe³⁺ has been observed for other Fe³⁺

346 oxides such as ferrihydrite and goethite (Leuz et al. 2006; Mitsunobu et al. 2010). Courtin-Nomade et
347 al. (2012) reported low concentrations of Sb in streams draining the mine waste dumps, and suggested
348 that this was due to the insolubility of this jarosite under the prevailing conditions (circum-neutral pH)
349 or to the formation of secondary Sb-bearing products from the dissolution of the Sb-bearing jarosite.

350 The reductive dissolution of Sb⁵⁺-bearing jarosite by Fe²⁺ under acidic pH (4.0 and 5.5) and
351 neutral pH (7) conditions has been examined (Karimian et al. 2018). No transformation of the jarosite,
352 nor release of Sb⁵⁺ was observed at pH 4.0 (Fig. 5). This was attributed to the surface of jarosite being
353 positively charged at pH 4.0, thus decreasing the possibility of Fe²⁺ to sorb, and hindering electron
354 transfer and dissolution. In contrast, at pH 5.5, the jarosite transformed to metastable green rust sulfate
355 within c. 60 min. and then to goethite within 24 h (Fig. 5). The pH used in these experiments was very
356 close to point of zero charge (5.6; Xu et al., 2013), and therefore the jarosite surface is near-neutrally
357 charged, allowing greater interaction of Fe²⁺ with the surface and formation of the green rust sulfate
358 and goethite. In these experiments, almost all of the released Sb⁵⁺ was structurally incorporated within
359 these secondary phases. At pH 7, the Sb was rapidly mobilized into the aqueous phase as Sb⁵⁺, but 75%
360 of this Sb⁵⁺ was structurally incorporated into secondary lepidocrocite (at Fe²⁺ = 1 mM), green rust and
361 goethite (at Fe²⁺ = 20 mM) (Fig. 5).

362 363 **IMPLICATIONS**

364 The incorporation of As in alunite, jarosite and beudantite group minerals (by replacement of
365 SO₄²⁻ by AsO₄³⁻ in the T-site of the mineral structure), and of Sb in jarosite group minerals (by
366 replacement of Fe³⁺ by Sb⁵⁺ in the B-site; Fig. 1) has important implications for storage of these
367 metalloid contaminants in acidic environments. Since incorporation of Pb²⁺ in the jarosite structure has
368 been shown to cause a higher degree of As take up than with no Pb²⁺ (Aguilar-Carrillo et al. 2018),
369 further research should be carried out to evaluate the effectiveness of this and other ions for

370 incorporation of greater amounts of As and Sb in alunite, jarosite and beudantite group minerals.
371 Overall, there is a paucity of information on the uptake of Sb in these minerals, despite the fact that it
372 has been shown to be incorporated in jarosite, and that As and Sb are often co-located in contaminated
373 environments (e.g., Craw et al. 2004). This research gap needs to be filled, as does further work on the
374 longer term (>1 year) stability, and abiotic and, especially, microbial dissolution, of As- and Sb-bearing
375 alunite, jarosite and beudantite group minerals.

376

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390 **REFERENCES**

391 Abdul, K.S.M., Jayasinghe, S.S., Chandana, E.P.S., Jayasumana, C., De Silva, P.M.C.S. (2015) Arsenic and
392 human health effects: A review. *Environmental Toxicology and Pharmacology*, 40, 828-846.

- 393 Acero, P., Ayora, C., Torrentó, C., and Nieto, J.-M. (2006) The behavior of trace elements during
394 schwertmannite precipitation and subsequent transformation into goethite and jarosite. *Geochimica et*
395 *Cosmochimica Acta*, 70, 4130-4139.
- 396 Aguilar-Carrillo, J., Villalobos, M., Pi-Puig, T., Escobar-Quiroz, I.N., and Romero, F.M. (2018) Synergistic
397 arsenic(V) and lead(II) retention on synthetic jarosite. I. Simultaneous structural incorporation behaviour and
398 mechanism. *Environmental Science: Processes & Impacts*, 20, 354-369.
- 399 Alpers, C.N., Rye, R.O., Nordstrom, D.K., White, L.D., and King, B. (1992) Chemical, crystallographic and
400 stable isotopic properties of alunite and jarosite from acid-hypersaline Australian lakes. *Chemical Geology*,
401 96, 203-226.
- 402 Becker U., and Gasharova B. (2001) AFM observations and simulations of jarosite growth at the molecular
403 scale: probing the basis for incorporation of foreign ions into jarosite as a storage mineral. *Physics and*
404 *Chemistry of Minerals*, 28, 545-556.
- 405 Bigham, J.M., and Nordstrom, D.K. (2000) Iron and aluminium hydroxysulfates from acid sulfate waters. In
406 Alpers, C.N., Jambor, J.L. & Nordstrom, D.K. (eds.) *Sulfate Minerals. Crystallography, Geochemistry and*
407 *Environmental Significance. Reviews in Mineralogy and Geochemistry*, 40, pp. 351-403.
- 408 Cooper, M.A., and Hawthorne, F.C. (2012) Refinement of the crystal structure of zoned philipsbornite-hidalgoite
409 from the Tsumeb mine, Namibia, and hydrogen bonding in the $D^{2+}G_3^{3+}(T^{5+}O_4)(TO_3OH)(OH)_6$ alunite
410 structures. *Mineralogical Magazine*, 76, 839-849.
- 411 Courtin-Nomade, A., Rakotoarisoa, O., Bril, H., Brybos, M., Forestier, L., Foucher, F., and Kunz, M. (2012)
412 Weathering of Sb-rich mining and smelting residues: Insight in solid speciation and soil bacteria toxicity.
413 *Chemie der Erde*, 72, 29-39.
- 414 Courtin-Nomade, A., Waltzing, T., Evrard, C., Soubrand, M., Lenain, J.-F., Ducloux, E., Ghorbel, S., Grosbois,
415 C., and Bril, H. (2016) Arsenic and lead mobility: From tailings materials to the aqueous compartment.
416 *Applied Geochemistry*, 64, 10-21.
- 417 Craw, D., Wilson, N., and Ashley, P.M. (2004) Geochemical controls on the environmental mobility of Sb and
418 As at mesothermal antimony and gold deposits. *Applied Earth Science*, 113, 3-10.

- 419 D'Angeli, I.M., Carbone, C., Nagostinis, M., Parise, M., Vattano, M. Madonia, G., and De Waele, J. (2018) New
420 insights on secondary minerals from Italian sulfuric acid caves. *International Journal of Speleology*, 47, 271-
421 291.
- 422 Dutrizac, J.E., and Jambor, J.L. (1987) The behaviour of arsenic during jarosite precipitation: Arsenic
423 precipitation at 97°C from sulphate or chloride media. *Canadian Metallurgical Quarterly*, 26, 91-101.
- 424 Dutrizac, J.E., and Jambor, J.L. (2000) Jarosites and their application in hydrometallurgy. In Alpers, C.N.,
425 Jambor, J.L. & Nordstrom, D.K. (eds.) *Sulfate Minerals. Crystallography, Geochemistry and Environmental*
426 *Significance. Reviews in Mineralogy and Geochemistry*, 40, pp. 405-452.
- 427 Egal, M., Casiot, C., Morin, G., Parmentier, M., Bruneel, O., Lebrun, S., and Elbaz-Poulichet, F. (2009) Kinetic
428 control on the formation of tooeleite, schwertmannite and jarosite by *Acidithiobacillus ferrooxidans* strains in
429 an As(III)-rich acid mine water. *Chemical Geology*, 265, 432-441
- 430 Feng, R., Wei, C., Tu, S., Ding, Y., Want, R., and Guo, J. (2013) The uptake and detoxification of antimony by
431 plants: A review. *Environmental and Experimental Botany*, 96, 28-34.
- 432 Filippi, M., Drahotka, P., Machvič, Böhmová, V., and Mihaljevič, M. (2015) Arsenic mineralogy and mobility in
433 the arsenic-rich historical mine waste dump. *Science of the Total Environment*, 536, 713-728.
- 434 Flores, M., Patiño, F., Palacios, E.G., Reyes, I., Reyes, M., Flores, V.H., Juárez, J.C., and Pandiyan, T. (2016)
435 The behaviour of arsenic during the thermal and chemical decomposition of the ammonium-arsenic jarosite.
436 Preprints, doi:10.20944/preprints201610.0059.v1
- 437 Forray, F.L., Smith, A.M.L., Navrotsky, A., Wright, K., Hudson-Edwards, K.A., and Dubbin, W.E. (2014)
438 Synthesis, characterization and thermochemistry of Pb-As-, Pb-Cu- and Pb-Zn-jarosite compounds.
439 *Geochimica et Cosmochimica Acta*, 127, 107-119.
- 440 Foster, A.L., Brown, G.E., Tingle, T.N., and Parks, G.A. (1998) Quantitative arsenic speciation in mine tailings
441 using X-ray absorption spectroscopy. *American Mineralogist*, 83, 553-568.
- 442 Gale, J.D., Wright, K. & Hudson-Edwards, K.A. (2010) A first principles determination of the orientation of
443 H_3O^+ in hydronium alunite. *American Mineralogist*, 95, 1109-1112.

- 444 Gieré, R., Sidenko, N.V., and Lazareva, E.V. (2003) The role of secondary minerals in controlling the migration
445 of arsenic and metals from high-sulfide wastes (Berikul gold mine, Siberia). *Applied Geochemistry*, 18,
446 1347-1359.
- 447 Gräfe, M., Beattie, D.A., Smith, E., Skinner, W.M., and Singh, B. (2008) Copper and arsenate co-sorption at the
448 mineral-water interfaces of goethite and jarosite. *Journal of Colloid and Interface Science*, 322, 399-413.
- 449 Hawthorne, F.C., Krivovichev, S.V., and Burns, P.C. (2000) The crystal chemistry of sulfate minerals. In Alpers,
450 C.N., Jambor, J.L. & Nordstrom, D.K. (eds.) *Sulfate Minerals. Crystallography, Geochemistry and*
451 *Environmental Significance. Reviews in Mineralogy and Geochemistry*, 40, pp. 1-112.
- 452 Henao, D.M.O., and Godoy, M.A.M. (2010) Jarosite pseudomorph formation from arsenopyrite oxidation using
453 *Acidithiobacillus ferrooxidans*. *Hydrometallurgy*, 104, 162-168.
- 454 Hudson-Edwards, K.A., Schell, C. and Macklin, M.G. (1999) Mineralogy and geochemistry of alluvium
455 contaminated by metal mining in the Rio Tinto area, southwest Spain. *Applied Geochemistry*, 14, 1015-1030.
- 456 Hudson-Edwards, K.A., Jamieson, H.E., Charnock, J.M., and Macklin, M.G. (2005) Arsenic speciation in waters
457 and sediment of ephemeral floodplain pools, Ríos Agrio-Guadiamar, Aznalcóllar, Spain. *Chemical Geology*,
458 219, 175-192.
- 459 Jambor, J.L. (1999) Nomenclature of the alunite supergroup. *Canadian Mineralogist*, 37, 1323-1341.
- 460 Jambor, J.L., Owens, A.R., Grice, J.D., and Feinglos, M.N. (1996) Gallobeudantite, $\text{PbGa}_3[(\text{AsO}_4)_2(\text{SO}_4)]_2(\text{OH})_6$,
461 a new mineral species from Tsumbe, Namibia, and associated new gallium analogues of the alunite-jarosite
462 family. *Canadian Mineralogist*, 34, 1305-1315.
- 463 Jamieson, H.E., Robinson, C., Alpers, C.N., Nordstrom, D.K., Poustovetov, A., and Lowers, H.A. (2005) The
464 composition of coexisting jarosite-group minerals and water from the Richmond mine, Iron Mountain,
465 California. *Canadian Mineralogist* 43, 1225-1242.
- 466 Johnston, S.G., Burton, E.D., Keene, A.F., Planer-Friedrich, B., Voegelin, A., Blackford, M.G., and Lumpkin,
467 G.R. (2012) Arsenic mobilization and iron transformations during sulfidization of As(V)-bearing jarosite.
468 *Chemical Geology* 334, 9-24.
- 469 Karimian, N., Johnston, S.G., and Burton, E.D. (2017) Antimony and arsenic behavior during Fe(II)-induced
470 transformation of jarosite. *Environmental Science & Technology*, 51, 4259-4268.

- 471 Karimian, N., Johnston, S.G., and Burton, E.D. (2018) Antimony and arsenic partitioning during Fe²⁺-induced
472 transformation of jarosite under acidic conditions. *Chemosphere*, 195, 515-523.
- 473 Kendall, M.R., Madden, A.S., Elwood Madden, M.E., and Hu, Q. (2013) Effects of arsenic incorporation on
474 jarosite dissolution rates and reaction products. *Geochimica et Cosmochimica Acta*, 112, 192-207.
- 475 Kolitsch, U., Slade, P.G., Tiekink, E.R.T., and Pring, A. (1999) The structure of antimonian dussertite and the
476 role of antimony in oxysalt minerals. *Mineralogical Magazine*, 63, 17-26.
- 477 Kocourková, E., Sracke, O., Houzar, S., Cempírek, J., Losos, Z., Filip, J., and Hřelová (2011) Geochemical and
478 mineralogical control on the mobility of arsenic in a waste rock pile at Dlouhá Ves, Czech Republic. *Journal*
479 *of Geochemical Exploration*, 110, 61-73.
- 480 Kubisz J. (1964) A study of minerals in the alunite–jarosite group. *Polska Akademia Nauk-Prace Museum*
481 *Ziemi*, 22, 1–93.
- 482 Kubisz J. (1970) Studies on synthetic alkali-hydronium jarosites: I. Synthesis of jarosite and natrojarosite.
483 *Mineralogia Polonica*, 1, 47–57.
- 484 Leuz, A.-K., Mönch, H., and Johnson, C.A. (2006) Sorption of Sb(III) and Sb(V) to goethite: influence on
485 Sb(III) oxidation and mobilization. *Environmental Science & Technology*, 40, 7277-7282.
- 486 Luo, Z.Q., Zhou, X.T., Jia, Q.M., Chen, X.F., Tao, Z.CH., and Liu, S.Q. (2015) Preparation of arsenical-
487 natroalunite solid solutions with high crystallinity by hydrothermal method. *Materials Research Innovations*,
488 19, No. Sup. 6, S6-S26.
- 489 Mitsunobu, S., Muramatsu, C., Watanabe, K., and Sakata, M. (2013) Behavior of antimony (V) during the
490 transformation of ferrihydrite and its environmental implications. *Environmental Science and Technology*,
491 47, 9660-9667.
- 492 Mitsunobu, S., Takahashi, Y., Terada, Y., and Skata, M. (2010) Antimony(V) incorporation into synthetic
493 ferrihydrite, goethite and natural iron oxyhydroxides. *Environmental Science & Technology*, 44, 3712-3718.
- 494 Nieto, J.M., Capitn, M.A., Sez, R., and Almodóvar, G.R. (2003) Beudantite: A natural sink for As and Pb in
495 sulphide oxidation processes. *Applied Earth Science B*, 112, 293-296.
- 496 Nordstrom, D.K. (1982) The effect of sulfate on aluminum concentrations in natural waters: some stability
497 relations in the system Al₂O₃-SO₂-H₂O at 298 K. *Geochimica et Cosmochimica Acta*, 46, 681-692.

- 498 Paktunc, D., and Dutrizac, J.E. (2003) Characterization of arsenate-for-sulfate substitution in synthetic jarosite
499 using X-ray diffraction and X-ray absorption spectroscopy. *Canadian Mineralogist*, 41, 905-919.
- 500 Paktunc, D., Foster, A., and Laflamme, G. (2003) Speciation and characterization of arsenic in Ketz River mine
501 tailings using X-ray absorption spectroscopy. *Environmental Science & Technology*, 37, 2067-2074.
- 502 Patiño, F., Flores, M.U., Reyes, I.A., Reyes, M., Hernández, J., Rivera, I., and Juárez, J. (2013) Alkaline
503 decomposition of synthetic jarosite with arsenic. *Geochemical Transactions*, 14:2.
- 504 Roca, A., Viñals, J., Arranz, M., and Calero, J. (1999) Characterization and alkaline decomposition/ cyanidation
505 of beudantite-jarosite materials from Rio Tinto gossan ores. *Canadian Metallurgical Quarterly*, 38, 93-103.
- 506 Roussel, C., Néel, C., and Bril, H. (2000) Minerals controlling arsenic and lead solubility in an abandoned gold
507 mine tailings. *Science of the Total Environment*, 263, 209-219.
- 508 Sánchez, L., Cruells, M., Roca, A. (1996) Sulfidization-cyanidation of jarosite species: applicability to the
509 gossan ores of Rio Tinto. *Hydrometallurgy*, 42, 35-49.
- 510 Savage, K.S., Bird, D.K., and Ashley, R.P. (2000) Legacy of the California Gold Rush: Environmental
511 geochemistry of arsenic in the southern Mother Lode Gold District. *International Geology Review*, 42, 385-
512 415.
- 513 Savage, K.S., Bird, D.K., and O-Day, P.A. (2005) Arsenic speciation in synthetic jarosite. *Chemical Geology*,
514 215, 473-498.
- 515 Smeaton, C.M., Walshe, G.E., Smith, A.M.L., Hudson-Edwards, K.A., Dubbin, W.E., Wright, K., Beale, A.M.,
516 Fryer, B.J., and Weisener, C.G. (2012) Simultaneous reduction of Fe and As during the reductive dissolution
517 of Pb-As jarosite by *Shewanella putrefaciens* CN32. *Environmental Science & Technology*, 46, 12823-
518 12831.
- 519 Smith, A.M.L., Dubbin, W.E., Wright, K., and Hudson-Edwards, K.A. (2006) Dissolution of lead- and lead-
520 arsenic jarosites at pH 2 and 8: insights from batch experiments. *Chemical Geology*, 229, 344-361.
- 521 Sundar, S., and Chakravarty, J. (2010) Antimony toxicity. *International Journal of Environmental Research and*
522 *Public Health*, 7, 4267-4277.
- 523 Sunyer, A., and Viñals, J. (2011a) Arsenate substitution in natroalunite: A potential medium for arsenic
524 immobilization. Part 1: Synthesis and compositions. *Hydrometallurgy*, 109, 54-64.

- 525 Sunyer, A., and Viñals, J. (2011b) Arsenate substitution in natroalunite: A potential medium for arsenic
526 immobilization. Part 2: Cell parameters and stability tests. *Hydrometallurgy* 109, 106-115.
- 527 Sunyer, A., Currubí, M., and Viñals, J. (2013) Arsenic immobilization as alunite-type phases: The arsenate
528 substitution in alunite and hydronium alunite. *Journal of Hazardous Materials*, 261, 559-569.
- 529 Szymanski, J.T. (1988) The crystal structure of beudantite, $Pb(Fe,Al)_3[(As,S)O_4](OH)_6$. *Canadian Mineralogist*,
530 26, 923-932.
- 531 Viñals, J., Sunyer, A., Molera, P., Cruells, M., and Llorca, N. (2010) Arsenic stabilization of calcium arsenate
532 waste by hydrothermal precipitation of arsenical natroalunite. *Hydrometallurgy*, 104, 247-259.
- 533 Welch, S.A., Christy, A.G., Kirste, D., Beavis, S.G. & Beavis, F. (2007) Jarosite dissolution I - Trace cation
534 flux in acid sulfate soils. *Chemical Geology*, 245, 183-197.
- 535 WHO (World Health Organization) (1996) Health criteria and other supporting information. Pp. 940-949 in:
536 *Guidelines for Drinking-water Quality*, 2nd edition, Vol. 2. WHO, Geneva.
- 537 WHO (World Health Organization) (1998) Addendum to Vol. 2. Pp. 281-283 in: *Guidelines for drinking- water*
538 *quality*, 2nd edition. WHO, Geneva.
- 539 Xu, Z., Lü, B., Wu, J., Zhou, L., and Lan, Y. (2013) Reduction of Cr(VI) facilitated by biogenetic jarosite and
540 analysis of its influencing factors with response surface methodology. *Materials Science and Engineering C*,
541 33, 3723-3729.
- 542 Zheng, M.X., Xu, J.M., Smith, L., and Naidu, R. (2003) Why a fern (*Pteris multifida*) dominantly growing on an
543 arsenic/heavy metal contaminated soil does not accumulate arsenic. *Journal de Physique IV*, 107, 1409-1411.
- 544